

1 **Histopathological changes in the Greenshell mussel, *Perna canaliculus*, in response to**
2 **chronic thermal stress**

3 Joanna S. Copedo^{1,2*}, Stephen C. Webb¹, Norman L. C. Ragg¹, Jessica A. Ericson¹, Leonie
4 Venter², Alfonso J. Schmidt³, Natalí J. Delorme¹, Andrea C. Alfaro²

5 ¹*Cawthron Institute, Private Bag 2, Nelson 7042, New Zealand.*

6
7 ²*Aquaculture Biotechnology Research Group, Department of Environmental Science, School*
8 *of Science, Auckland University of Technology, Private Bag 92006, Auckland 1142, New*
9 *Zealand.*

10 ³*Hugh Green Cytometry Centre, Malaghan Institute of Medical Research, PO Box 7060,*
11 *Wellington 6242, New Zealand*

12 *Corresponding author: Joanna.copedo@cawthron.org.nz

13 **Abstract**

14 Climate change associated temperature challenges pose a serious threat to the marine
15 environment. Elevations in average sea surface temperatures are occurring and increasing
16 frequency of marine heatwaves resulting in mortalities of organisms are being reported. In
17 recent years, marine farmers have reported summer mass mortality events of the New Zealand
18 Greenshell mussel, *Perna canaliculus*, during the summer months; however, the etiological
19 agents have yet to be determined. To elucidate the role of thermal stress, adult *P. canaliculus*
20 were exposed to three chronic temperature treatments: a benign control of 17°C and stressful
21 elevations of 21°C and 24°C. Eight mussels per treatment were collected each month
22 throughout a 14-month challenge period to identify and investigate histopathological
23 differences among *P. canaliculus* populations exposed to the three temperatures.
24 Histopathology revealed several significant deleterious alterations to tissues associated with
25 temperature and exposure time. Increasing temperature and progression of time resulted in 1)
26 an increase in the number of focal lipofuscin-ceroid aggregations, 2) an increase in focal
27 hemocytosis, 3) an increase in the thickness of the sub-epithelial layer of the intestinal tract
28 and 4) a decreased energy reserve cell (glycogen) coverage in the mantle. Prolonged exposure,
29 irrespective of temperature, impacted gametogenesis, which was effectively arrested.
30 Furthermore, increased levels of the heat shock protein 70 kDa (HSP 70) were seen in gill and
31 gonad from thermally challenged mussels. The occurrence of the parasite *Perkinsus olseni* at
32 month 5 in the 24°C treatment, and month 7 at 21°C was unexpected and may have exacerbated
33 the fore-mentioned tissue conditions. Prolonged exposure to stable thermal conditions

34 therefore appears to impact *P. canaliculus*, tissues with implications for broodstock captivity.
35 Mussels experiencing elevated, temperatures of 21 and 24°C demonstrated more rapid
36 pathological signs. This research provides further insight into the complex host-pathogen-
37 environment interactions for *P. canaliculus* in response to prolonged elevated temperature.

38 **Key words**

39 Environmental stress, Greenshell mussel, histopathology, gametogenesis, atresia, *Perkinsus*
40 *olseni*

41

42

43

44 **1. Introduction**

45 Climate observations have demonstrated the role of climate change on global environmental
46 stressors for more than a decade (Thomas et al., 2004; Mooney et al., 2009). Oceanic
47 environmental challenges as a result of climate change include wider fluctuations in pH,
48 salinity and temperature (Boyd et al., 2014). Not only are elevated average sea temperatures
49 expected (Houghton et al., 2001), but marine heatwaves (MHW) are likely to increase in
50 frequency, particularly in coastal regions (Lubchenco et al., 1993; Houghton et al., 2001; Petes
51 et al., 2007; Filgueira et al., 2016; Smale et al., 2019). Ocean temperatures have already
52 increased by an average of 1.5°C over the past century and will continue rising (Hobday et al.,
53 2016; Tuckett et al., 2017). Marine heatwaves are of particular interest due to correlations
54 between elevated temperature and mass die-offs of marine organisms during the summer
55 period, loosely termed “summer mortalities” (Salinger et al., 2020a). Heatwave-related
56 mortalities have impacted several molluscan populations and aquaculture industries
57 worldwide, with several environmental stressors, including high temperature and pathogens
58 (e.g., *Perkinsus marinus*) being identified as causal agents (Harvell et al., 1999; Garrabou et
59 al., 2009; Rubio-Portillo et al., 2016).

60

61 New Zealand’s coastal waters host a rich and diverse ecosystem, which is greatly impacted by
62 temperature elevations (Stenton-Dozey et al., 2020; Behrens et al., 2022). In addition to long-
63 term warming of New Zealand’s coastal waters, there has been a rise in MHW frequency
64 (Oliver et al., 2017; Salinger et al., 2019; Sutton and Bowen, 2019; Behrens et al., 2022). An

65 unprecedented MHW in the austral summer of 2017/2018 resulted in a mass mortality event of
66 bull kelp (*Durvillea* spp.) in the South Island of New Zealand (Thomsen et al., 2019; Salinger
67 et al., 2020b). As a result of increasing frequency of MHW events, it is likely that sea surface
68 temperature (SST) will exceed thermal optima for local species, resulting in increased
69 detrimental thermal exposure for many marine organisms. Elevated temperatures can
70 significantly affect processes such as, metabolism, reproduction, immune response, and growth
71 (Portner, 2002; Angilletta Jr and Angilletta, 2009; Filgueira et al., 2016; Dunphy et al., 2018).
72 Survival of thermal stress requires a fine balance between cellular response, cell integrity and
73 homeostatic capacity. Furthermore, stress compromises the immune system resulting in disease
74 susceptibility (Trump et al., 1997; Manduzio et al., 2005; Fulda et al., 2010; Carella, 2015;
75 Delorme et al., 2021). When an animal is exposed to sublethal stressors such as increased
76 temperature, reproduction can also be inhibited as energy is reallocated to defence and tissue
77 repair (Michalek-Wagner and Willis, 2001; Petes et al., 2007). In addition, timing of
78 reproduction can be disrupted and result in asynchronicity of spawning and decreased
79 fertilization (Walther et al., 2002; Philippart et al., 2003; Petes et al., 2007). Response to these
80 increasing heatwave events is, therefore, likely to lead to various biological consequences and
81 may even alter phenology and species ranges (Parmesan and Yohe, 2003; Petes et al., 2007;
82 Seuront et al., 2019; Delorme et al., 2021).

83

84 The Greenshell mussel, *Perna canaliculus* (Gmelin 1791), is endemic to New Zealand (Jefferies
85 et al., 1999; Alfaro et al., 2001) naturally inhabiting a wide range of temperatures and salinities
86 (Jefferies et al., 1999). It is not only an ecologically and culturally valuable species, but *P.*
87 *canaliculus* farming is considered the cornerstone of the aquaculture industry in New Zealand
88 (Dawber et al., 2004; Stenton-Dozey et al., 2020). Unlike more mobile species such as fish,
89 mussels have a limited capacity to relocate when conditions, such as increasing seawater
90 temperatures, become unfavourable. Due to this sedentary lifestyle mussels, such as the *P.*
91 *canaliculus*, have developed effective strategies to minimise the impact when exposed to short
92 term environmental stressors (e.g. de la Ballina et al., 2022). Key strategies include the ability
93 to close their shells thereby reducing contact to the soft tissues, and physiological adaptations
94 such as the development of an effective immune system (e.g. Gosling, 2008; Chong, 2022; de
95 la Ballina et al., 2022). The importance of *P. canaliculus* has prompted various research
96 projects in relation to thermal stress (e.g. Ren and Ross, 2005; Ren et al., 2020; Stenton-Dozey
97 et al., 2020). Effects of acute thermal stressors on *P. canaliculus* have previously been studied

98 to some extent (e.g. Dunphy et al., 2013; Dunphy et al., 2015; Dunphy et al., 2018; Delorme et
99 al., 2020a; Delorme et al., 2021); however, very few studies have focused on the effects of
100 prolonged thermal stress (Ericson et al., 2023). Previous works include recent metabolomic
101 research on *P. canaliculus* during a summer heatwave event and acute, and chronic, heat
102 exposure studies. These have found signs of stress and disruptions to metabolic pathways
103 relating to, for example, energy metabolism and oxidative stress (Li et al., 2020; Nguyen and
104 Alfaro, 2020; Delorme et al., 2021; Ericson et al., 2023). Metabolomic assessments provide a
105 snapshot of the biological response at the time of sampling (Alfaro and Young, 2018), in
106 contrast histology can give insights into cumulative effects on tissues that may develop over
107 time (Costa, 2018). Consequently, this method may provide further insights into the effects of
108 prolonged thermal stress and marine heatwave exposure at the tissue level of *P. canaliculus* in
109 New Zealand as these are largely unknown, as are the etiological agents of thermally mediated
110 mortalities.

111

112 Histopathology is an essential tool in the presumptive diagnosis of diseases and mortality
113 investigations as part of health surveillance (Hooper et al., 2014; Knowles et al., 2014; Costa,
114 2018). It is a bench-mark screening method that can provide a phenotypic anchor point for
115 specific data as it provides tissue-level measures of general, reproductive, and metabolic
116 condition as well as pathogen detection (Bignell et al., 2008; Zannella et al., 2017). In the
117 present study, histopathology, and immunohistochemistry (IHC) were used to elucidate
118 alterations in *P. canaliculus* tissues, as well as presence and prevalence of pathogens, when
119 exposed to prolonged thermal stress. Temperatures were selected based on a summer ambient
120 (17°C SST), present-day high summer temperatures (21°C high summer SST), and a projected
121 high according to a +2.5°C predicted increase (24°C projected summer SST by the year 2100)
122 (Law et al., 2017; Ericson et al., 2023). Identifying alterations to *P. canaliculus* tissue condition
123 and constitution in relation to increasing and prolonged temperature exposure is required to
124 observe the effects of stressors on the tissue structure and architecture. As such this research
125 aims to improve our knowledge on detecting thermal stress and correlation with tissue
126 condition and pathogens. Therefore, providing insights into potential impacts and implications
127 in response to future climate stressors.

128

129 **2. Methods**

130 **2.1. Experimental design**

131 Adult *Perna canaliculus* (80 – 120mm shell length, n=1215) were transported from a farm in
132 the Pelorus Sound, Marlborough, New Zealand to laboratories at the Cawthron Aquaculture
133 Park, Nelson, in May 2018. The mussels were weighed (± 1 g), shell length measured (± 1 mm)
134 and engraved with identification numbers (Dremel 3000 rotary tool with a 1.6mm round
135 diamond tip). The mussels were then distributed to nine well aerated 100 L tanks (n = 135 per
136 tank) designed for broodstock holding. After 6 weeks of acclimation to the current ambient
137 seawater ($16^{\circ}\text{C} \pm 2^{\circ}\text{C}$), the temperature was increased approximately 0.25° and 1.8°C per day
138 over two weeks depending on temperature treatment until the tanks reached the desired
139 temperatures: ‘benign’ 17°C , and elevated 21°C and 24°C , with three replicate tanks per
140 temperature (see Figure 1 in Ericson et al. (2023) for the temperature trajectory for each
141 treatment). The control temperature was maintained in three primary header tanks using a heat
142 pump set to 17°C , while the 21°C and 24°C treatment temperatures were obtained by heating
143 the control temperature using titanium and glass submersible heaters. Mussels were
144 continuously fed bloomed algae pond seawater throughout the trial supplemented with a 50:50
145 combination of monocultured *Tisochrysis lutea* (formerly *Isochrysis*
146 *galbana*) and *Chaetoceros calcitrans* (250L/day) for a full description of feed and water
147 conditions see Ericson et al. (2023). *P. canaliculus* individuals were weighed (g) and measured
148 (mm) after 2-, 5-, and 10-months exposure. Mortalities were reported daily and immediately
149 removed from the system. The shell lengths of dead animals were measured and recorded to
150 provide an indication of growth to mortality. Weighing was not performed due to loss of tissue
151 mass through rapid decay. This rapid decay also prevented harvest of dead specimens for
152 histopathology.

153 **2.2. Water quality sampling**

154 Seawater flows (2.5 ± 0.5 L/min) were measured twice weekly and adjusted as required. Water
155 quality parameters (dissolved oxygen, temperature, pH and chlorophyll-*a*) were measured
156 throughout the trial and reported within Ericson et al. (2023).

157 **2.3. Histopathology**

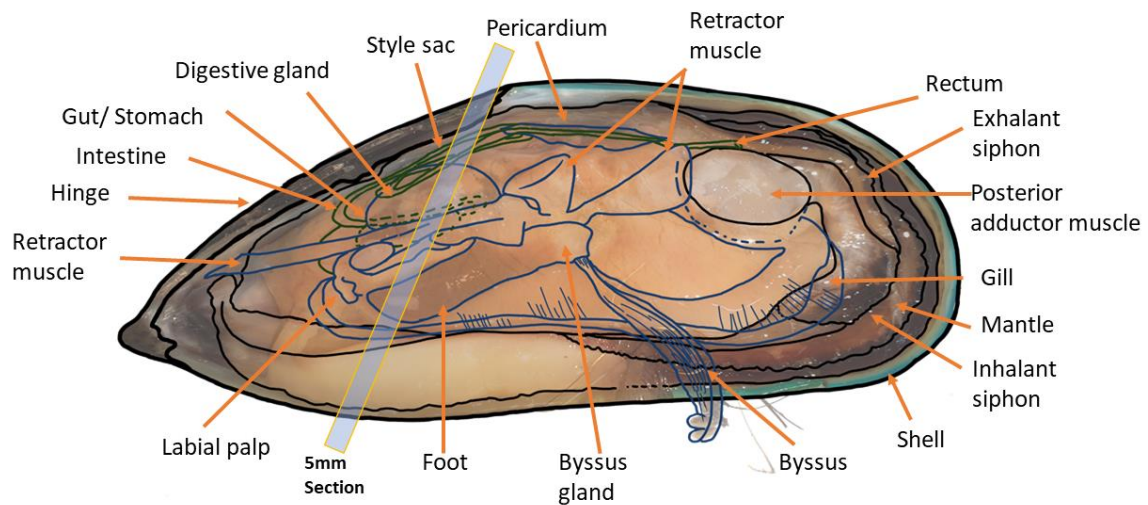
158 Eight *P. canaliculus* per temperature treatment (17°C , 21°C and 24°C) were randomly
159 collected across replicate tanks each month and prepared for histopathological assessment. An
160 additional 10 mussels per temperature, were collected at 4 time points throughout the trial
161 (labelled as month and letter B in Table. 1).

162 Table 1. Sampling numbers and collection months for each of the temperature treatments over
 163 the 15 months (July 2018 – Sept 2019). A fixation issue at month 3 resulted in a lack of
 164 individuals for the histology assessment. Thus, month 3 is represented graphically but was not
 165 included in the statistical analysis. Month 2B was the initial sampling, where all mussels were
 166 at 17°C, prior to raising the mussels to the desired temperature achieved by month 3 (None
 167 available (NA)). At month 15B there were no mussels left in the 24C treatment to sample (NA).

| | | Sampling month and date | | | | | | | | | | | | | | | |
|-------------|------|-------------------------|------------|-----------|-----------|-----------|-----------|-----------|-----------|-------------|-----------|-----------|-----------|-----------|-----------|------------|-----------|
| | | 2B | 3 | 4 | 5 | 6 | 6B | 7 | 8 | 9 | 10 | 11 | 11B | 12 | 13 | 14 | 15B |
| | | 2018 | | | | | | 2019 | | | | | | | | | |
| | | 21 Aug | 24 Sept | 19 Oct | 19 Nov | 01 Dec | 06 Dec | 02 Jan | 25 Feb | 03 April | 02 May | 04 Jun | 11 Jun | 10 Jul | 05 Aug | 30 Sept | 15 Oct |
| Temperature | 17°C | 15 | 1 | 8 | 8 | 8 | 10 | 8 | 8 | 8 | 8 | 8 | 10 | 8 | 8 | 8 | 18 |
| | 21°C | NA | 3 | 8 | 8 | 8 | 10 | 8 | 8 | 8 | 8 | 8 | 10 | 8 | 8 | 8 | 18 |
| | 24°C | NA | 0 | 8 | 8 | 8 | 10 | 8 | 8 | 8 | 8 | 8 | 10 | 8 | 8 | 4 | NA |

168

169 Tissues were removed intact from the shell and sectioned as shown in figure 1. The 5mm
 170 sectioned samples (mantle, gill, kidney, digestive gland, mid-gut, muscle, and nervous tissue)
 171 were placed into histological cassettes and fixed in a 4% formalin solution (1:9 v/v, 37%
 172 formaldehyde: 0.35µm filtered seawater). The samples were left in the 4% formalin solution
 173 for 48 hours before being transferred to a 70% ethanol solution (Howard, 2004). Samples were
 174 embedded in paraffin wax, sectioned (3-5µm) using a rotary microtome and stained using
 175 routine hematoxylin and eosin (H&E) (Howard, 2004).



176

177 Figure 1. Greenshell mussel (*Perna canaliculus*) general anatomy depiction. Rectangular
 178 window indicates the location of the 5mm histology section. This section was selected to
 179 maximise chances of acquiring all tissue types, including midgut, digestive gland, gill, gonad
 180 (mantle) and muscle.

181 **2.3.1. Tissue Analysis**

182 Histological observation under a light microscope (Olympus BX40), at a magnification of x40
 183 to x1000, was conducted on a range of tissues, assessing tissue-specific alterations by scoring
 184 semi-quantitatively and quantitatively as explained below in sections 2.3.2 to 2.3.4.

185 **2.3.1.1. Gill and digestive gland scoring**

186 The gill and the digestive gland were assessed at x200 and x400 magnification, then a score of
 187 tissue quality was devised based on structural deviations from those seen in mussels sampled
 188 on arrival (Baseline: month 2). The gill and digestive gland grading criteria were generalised
 189 and adapted from (Kim et al., 2006) and (Perez-Cebrecos et al., 2022), whereby 0= normal
 190 tissue and no obvious changes, 1= ‘mild’: minor alterations, 2= ‘moderate’: up to half of the
 191 organ affected and 3= ‘severe’ alterations with marked disruption to tissue architecture. Tissues
 192 were graded at each timepoint between the three temperatures (17°C, 21°C and 24°C) and then
 193 compared across time (16 months).

194

195 **2.3.1.2. Mantle and connective tissue scoring**

196 The mantle tissues were graded using four subjective semi-quantitative criteria, each graded
 197 from least to most severe: 1) the level of ceroid material found in the section (implicating
 198 oxidative stress or previous immune response, for this research ceroid and lipofuscin are
 199 considered together), 2) the level of focal hemocytosis (nodule or inflammatory capsule),
 200 indicative of an immune response (Table 2; Fig. 4), 3) the level of storage cell material in the
 201 mantle, and 4) the width of the subepithelial layer between the gut and the digestive tubule
 202 tissue (Table 3; Fig. 5). These scores were developed to assist with the identification of a stress
 203 or an immune response to increasing temperature, as well as to help standardise interpretation.
 204 The average scores of each of the four criteria for each temperature were plotted against time
 205 in a heat map table to detect changes in each related to their respective criteria.

206 Table 2. Semi-quantitative scoring of level of ceroid material and focal hemocytosis from least
 207 severe (1) to most severe (4). Criteria are similar to the grading scale of Carella et al., (2015)
 208 and Bignell et al. (2008) and specific illustrations are located in the results section..

| Score | Ceroid level criterion | Focal hemocytosis level criterion |
|--------------|---|--|
| 0 | No ceroid material observed | No hemocytosis alterations observed |
| 1 | Low level ceroid, focal alterations observed across tissues | Low level focal hemocytosis alterations observed across tissues |
| 2 | Moderate level of multifocal ceroid alterations observed across tissues | Moderate level multifocal hemocytosis alterations in multiple tissues |
| 3 | Severe level of multifocal ceroid alterations across tissues | Severe multifocal and hemocytosis alterations |
| 4 | Extensive diffuse ceroid accumulation in multiple tissue types | Extensive hemocytosis diffusely distributed across multiple tissue types |

209

210

211 Table 3. Semi-quantitative scoring of level of storage cell material (glycogen) and gut
 212 subepithelial thickening from least severe (1) to most severe (3). A score of 3 for storage cell
 213 coverage indicates glycogen rich mantle tissue whereas a score of 1 for the gut is indicative of
 214 healthy gut performance, with low numbers of hemocytes around the gut epithelial layer.

| Score | Storage cell (Glycogen) coverage criterion | Gut subepithelial layer extent criterion |
|--------------|--|--|
| 0 | No storage cells observed | None observed (no hemocytes observed around the epithelium). |
| 1 | Light coverage of storage granules in the mantle tissue. | Normal thickness the width of the gut epithelium \leq width of the epithelium. |

| | | |
|---|---|--|
| 2 | Moderate coverage of storage granules in the mantle tissue. | Thickness greater than thickness of the gut epithelial layer $\leq 2x$ thickness of epithelium. |
| 3 | Full coverage of storage granules in the mantle tissue | Thickness greatly increased expanding into the digestive tissue greater than thickness of gut epithelial layer (Abnormal). |

215

216

217 **2.3.1.3. Immunohistochemical detection of HSP70 in gill and gonad: A preliminary**
 218 **investigation**

219 Three paraffin blocks prepared from mussel tissues before temperature exposure (month 2b)
 220 and 3 blocks after temperature acclimation (month 4) were selected for preliminary assessment
 221 of HSP70 as a potential heat stress marker. A 4 μ m section was obtained from the paraffin
 222 blocks as processed and described in section 2.3. The paraffin section was then mounted on to
 223 a microscope glass slide. This section was then dewaxed, hydrated, and processed through a
 224 heat-mediated antigen retrieval procedure for 5 minutes at 120°C. The prepared slides were
 225 stained using an indirect antibody approach with a primary monoclonal antibody reactive to
 226 HSP70 (MA3-006, Invitrogen) used at a concentration of 1:200. A complementary secondary
 227 antibody conjugated with Alexa fluor 647 (A32728, Invitrogen) was used at a concentration of
 228 1:1000. DAPI (0.2 μ g ml⁻¹) was then used as a counter stain to highlight the cellular nuclei.
 229 Sections were mounted using an aqueous mounting media and 22x50 mm (1.5 H) cover slip.

230 Slides were observed under an inverted microscope (IX83, Olympus, Tokyo, Japan) equipped
 231 with a laser scanning confocal head (FV3000, Olympus, Tokyo, Japan), using a 405nm (50
 232 mW) and a 640nm (40mW) laser line, and Fluoview FV31S-SW software (version 2.3.2,
 233 Olympus Corp.). Images of gill and gonad tissue were acquired using the UplanSApo 20x N.A.
 234 0.75 objective and a confocal aperture of 125 μ m. These acquisition conditions were
 235 established according to Nyquist sampling criterion and acquired signals that fulfil the Rose
 236 criterion for signal-to-noise ratio in all the marker channels (Pawley, 2006). Images were
 237 collected using the following emission filter configuration for each fluorophore: DAPI (430-
 238 470nm), AF647 (650-700nm). Image visualisation and formatting was performed using ImageJ
 239 (Schindelin et al., 2012).

240 **2.3.1.4. Reproductive staging**

241 Gonad development scoring proceeded using the following system similar to Kennedy (2010),
242 Alfaro et al. (2001) and Buchanan (2001) (Table 4). Stage definitions were altered to include
243 maturity of the oocytes as well as mantle coverage. Females and males were scored based on
244 the criteria of stages and given a numerical gonad index value using microscopy (Table. 4).
245 The frequency or proportion at each stage was plotted for visual representation. The gonadal
246 index (GI) was then calculated based on formulae by King et al. (2009), Galinou-Mitsoudi and
247 Sinis (1994), Buchanan (2001) and Alfaro et al. (2001), whereby GI (Gonad index) is
248 determined by the average rank, or score, of the population. A GI of 0 is indicative that the
249 population is closer to the resting phase and a GI approaching 3 the population is close to full
250 maturity. Percent follicle coverage (FC) within the mantle tissue was estimated by eye followed
251 by image analysis. Area analysis was conducted on the 25 females identified within this trial
252 using cellSens™ software (Olympus cellsens Standard 3.1 [build 21199] on an Olympus BX53
253 compound) to validate the estimates. To validate the percent follicle coverage the area of the
254 frame was recorded then a free-form polygon was used to define the cross-sectional area of
255 each follicle contained within the frame. The percent area coverage was then calculated. The
256 values were then fitted to a regression and compared with the estimates. Estimates by eye were
257 developed to be used for routine determination in a commercial aquaculture setting where
258 image analysis is not always possible.

259 Table 4. Histological criteria of seven stages and the associated gonad index score for the
260 reproductive condition of *Perna canaliculus*.

| Stage | Gonad index score | Description | Male | Female |
|---------------|-------------------|--|--|---|
| Resting | 0 | Sex indeterminable. Resting or inactive some basal cells and/or very early precursors. Good coverage of storage reserve cells in mantle (glycogen). | | |
| Early | 1 | Gametogenesis has begun and small germinal cells are present. Follicles 500µm or less in diameter. | Spermatogonia inside follicles along the wall. Some spermatozoa are visible. | Oogonia line follicle wall, oocyte nuclei large attached to wall by cytoplasmic stalk |
| Late | 2 | Follicles become larger in size; 50% of follicles equal to, or larger, than medium size and/or both sexes have the potential to spawn and produce 50% or less of their mature sperm when force spawned. Follicles 500µm or more in diameter. | Concentric layers of spermatogonia, spermatocytes and spermatids. 50:50 Precursor to mature cell ratio. | Oocytes accumulate yolk and some mature oocytes are free and/or 50% distribution of precursor cells to mature oocytes. |
| Ripe | 3 | Mantle nearly full, with at least 80% coverage of the mantle, follicles are large (1000µm or more) and/or development of gonad is as follows for males and females. | Mantle filled and follicles show <10% precursor cells and with up to 50% of follicles with mature sperm converging in the centre but still dense. | Most oocytes at maximum size and mostly free from the wall, some small loss of oocytes. |
| Spawning | 2 | Release of gametes has begun, spillage of loose gametes evident upon histological sectioning. Evidence of a reduction in gamete density within follicle. Less than 80% atresia observed. | Dense band of ripe spermatozoa; greater than 80% spermatozoa converging in the centre with some follicles showing a loss in density. | Ripe oocytes present with a reduction in density of gametes in the centre of the follicle. Very few precursor cells (<10%). |
| Redevelopment | 1 | Follicles reduced in size and do not take up all the mantle area, remaining oocytes arranged loosely with new oocytes visible. | Dense band of ripe spermatids give rise to new lamellae, with gaps still present; very few follicles with phagocytosis developing (less than 50%). | Some atresic debris as well as phagocytes, but phagocytes are not affecting the new oocytes. |
| Spent | 0 | Follicles either empty or collapsing and degenerate amoebocytes/ phagocytes attack unspawned material, autolysis. Lumen filled with debris, atresia, and phagocytosis. Residual gametes will be broken down. | Phagocytes attack unspawned sperm. | Phagocytes clearing debris and follicles display greater than 80% oocyte atresia. |
| Atresia | | Autolysis, degradation, and reabsorption of oocytes. Includes cytoplasmic discolouration and darker staining, irregular jigsaw shape, retraction, and detachment of the cell membrane. Lysing of the cell membrane, oocyte contents spill into follicle. Can occur at any stage. | | |
| Phagocytosis | | Phagocytes clear unspawned gametes: the follicle will have cellular debris with some ceroid material visible. Can occur at any stage. | | |

261 Preliminary investigations of female gonad status and indications of presence of atresic oocytes
262 (defined in Table. 4) resulted in further analysis of the oocyte condition. Three micrographs
263 were taken of the gonad and mantle tissue for each female at 20x using an Olympus BX53
264 compound microscope and cellSens imaging software (Olympus cellsens Standard 3.1 [build
265 21199]). Oocytes in each of the micrographs were then quantitatively analysed using methods
266 by Chérel and Beninger (2017). Quantification of the oocytes was performed using
267 stereological counts and oocytes allocated to 3 groups, immature (I) (attached to follicle
268 (acinal) wall; stalked), mature (M) (detached from follicle wall) and atresic (A) (Fig. 2). The
269 three counts were then averaged and recorded to produce a single value per category for each
270 female and the importance of atresia was calculated using the following by Chérel and
271 Beninger (2017):

272 1) Percent atresia (PA) = $\frac{A}{I+M+A} * 100$

273

274 2) Percent mature (PM) = $\frac{M}{I+M+A} * 100$

275

276

277 3) Percent immature (PI) = $\frac{I}{I+M+A} * 100$

278

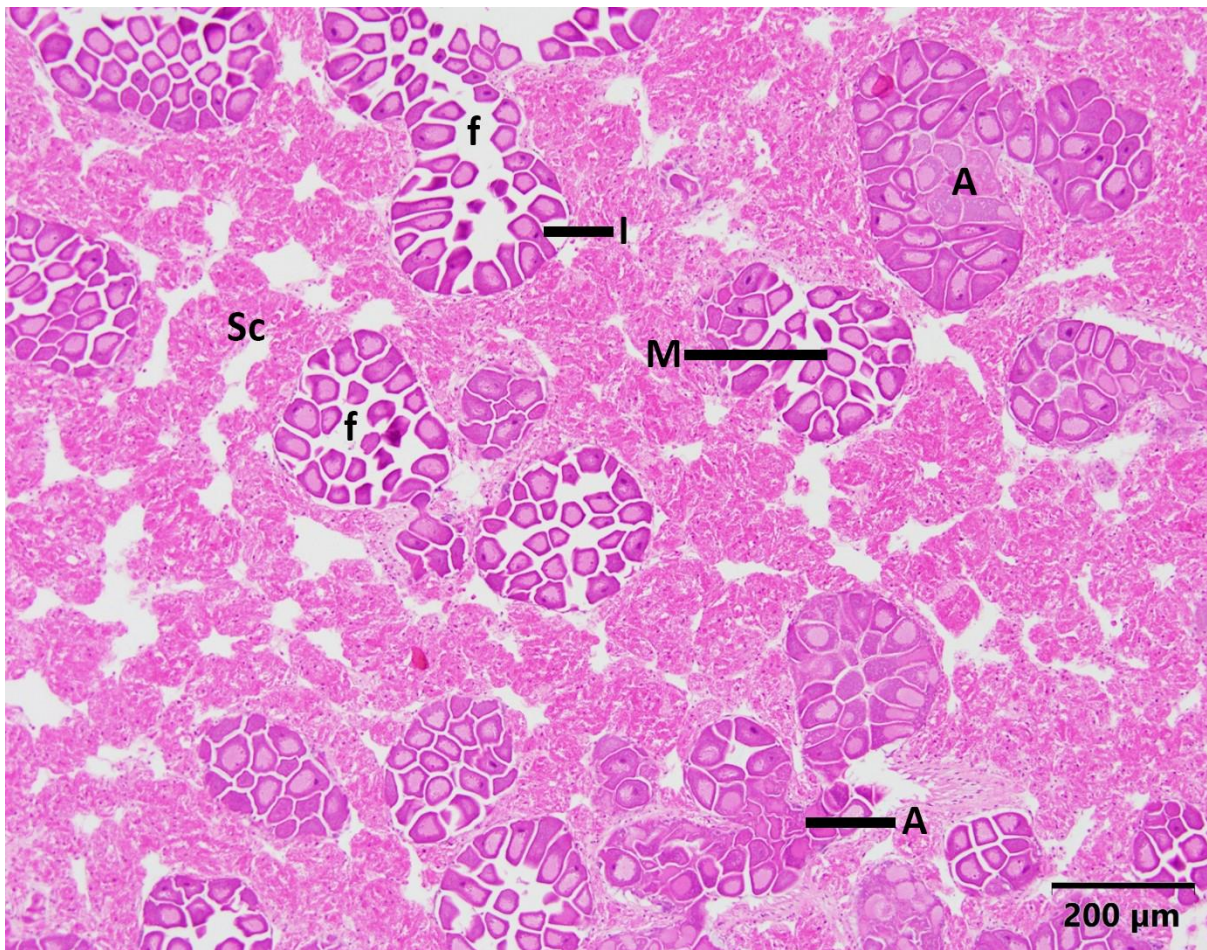
279

280 4) Atresic impact (AI) = $\frac{PA}{PA+PM} * 100$

281

282

283 The atresic impact was determined to be the minimum impact of atresia on the oocytes whereby
284 the fate was known (i.e., spawned as healthy oocytes and fate known). Health of immature
285 oocytes was yet to be determined and could either develop into healthy oocytes or become
286 atresic (Chérel and Beninger, 2017).



287

288 Figure 2. Female gonad stained with H&E indicating stages of oocytes immature (I), mature
 289 (M), atresic (A) within the follicles (f) and glycogen rich mantle storage cell (Sc).

290 **2.3.2.3. Pathogens**

291 Presence/absence of parasites was recorded and presented as proportion of population
 292 prevalence. The apicomplexan-X (APX) and *Perkinsus olseni* were the most prevalent and, as
 293 such, were investigated further (Fig. 5). The intensity of APX and *P. olseni* were scored using
 294 adapted methods from Hine (2002), Kim et al., (2006) and Suong (2018) (Table 5). The section
 295 was screened at x200 to identify cells, x400 and x1000 magnifications were then used for
 296 further identification. Following the histological identification of *P. olseni*, a selection of
 297 processed blocks was sent to the NZ Ministry for Primary Industry National Animal Health
 298 Laboratory and subjected to PCR screening using standard and specific primers to confirm
 299 diagnosis.

300 Table 5. Two parasites, APX and *Perkinsus olseni*, were scored using a 0 – 5 grading scale,
 301 where 0 indicated no parasites were observed and 5 is a severe level of parasites was observed
 302 (Hine, 2002; Kim et al., 2006; Suong et al., 2018).

| Score | APX: Criterion | <i>Perkinsus olsenii</i> : Criterion |
|-------|---|--|
| 0 | No parasites were observed in the tissue section. | No parasites were observed in the cross-section. |
| 1 | Low: Parasite difficult to detect <10 APX zoites were present in the tissue section after extensive search. | Low: <10 <i>Perkinsus</i> cells observed in tissues after extensive searching. For the purpose of this scoring rosette- stage groups were considered as 1. |
| 2 | Moderate: APX zoites were present in the tissue section and detected in small groups. Occasionally associated with ceroid material. | Moderate: <i>Perkinsus</i> cells were observed in tissues and more easily identified. |
| 3 | High: APX zoites were more easily detected, developing multiple moderate to large lesions of APX zoites and in most cases associated with ceroid. | High: <i>Perkinsus</i> cells observed to be distributed in multiple tissues <i>Perkinsus</i> cells. No lesions observed. |
| 4 | Heavy: APX zoites abundant in most tissues developing large lesions. | Heavy: <i>Perkinsus</i> cells abundant in most tissues, developing lesions. |
| 5 | Severe: APX zoites abundant with high tissue congestion and connective tissue destruction. | Severe: <i>Perkinsus</i> cells abundant in most tissues, developing large and abundant lesions. |

303

304 2.4. Statistical analysis

305 Statistical analyses were conducted using R version 4.0.3 (R Core Team, 2021) with the R
306 studio interface (RStudio Team, 2021). Histology data were analysed using both temperature
307 and time as explanatory factors. Quantitative data (follicle coverage and atresia) were checked
308 for normality and homogeneity of variance using the Shapiro-Wilks and Levene's tests,
309 followed by analysis using ANOVA. The atresia data were non-normal and unbalanced due to
310 the lack of female representation in each category. However, ANOVA Type II sum of squares
311 at the 0.05 level was used as the data met the requirement for homogeneity of variance (Quinn
312 and Keough, 2012). Analysis of the semi-quantitative data (hemocytosis, ceroid material,
313 mantle reserve cell, gut subepithelium and gonad index) was performed using a Generalised
314 linear mixed effect model in the MCMCglmm package (Hadfield, 2010). Tank number was
315 included as a random effect for the MCMCglmm however was tested and excluded from the
316 other analyses. Binomial general linear models were performed on the pathogen (APX and *P.*
317 *olsenii*) prevalence data. Ridge penalizers were added prior to the estimates to increase data
318 'sensitivity' if the differences between the responses were deemed as "perfect" (Cule and
319 Frankowski, 2021). A *p* value < 0.05 was considered statistically significant.

320 3. Results

321 3.1. Water quality and survival

322 On initiation of the trial, once seawater had reached the desired temperature, the treatments
 323 were maintained at that temperature. A logistical issue with the temperature system occurred
 324 in the 21°C and 24°C treatments during month 10 of the experiment and the temperature
 325 dropped to 17°C for 2 weeks (see Ericson et al., 2023). Analysis of mussel tissue alternations
 326 did not reveal obvious effects from this technical difficulty. Temperatures were slowly
 327 increased back to 21°C and 24°C once the system's malfunctioning was repaired. The dissolved
 328 oxygen and pH levels remained stable through the trial (Table 6).

329

330 **Table 6.** Water quality parameters (mean ± standard error (SE)) measured throughout the
 331 experiment.

| Water parameter | Desired temperature | | |
|------------------------|---------------------|--------------|--------------|
| | 17°C | 21°C | 24°C |
| Temperature (recorded) | 17.9 ± 0.5°C | 21.2 ± 1.0°C | 23.2 ± 1.6°C |
| Dissolved oxygen | 92.6 ± 1.3% | 90.5 ± 0.3% | 90.6 ± 1.3% |
| pH | 8.14 ± 0.01 | 8.12 ± 0.01 | 8.13 ± 0.01 |

332

333 Survival was significantly impacted by chronic temperature exposure to 24°C, and a mortality
 334 began at month 5 of the trial until a 100% mortality was reached at month 15. At 17°C and
 335 21°C, survival at completion of the trial was at 94% and 90%, respectively (see complementary
 336 study presented by Ericson et al. (2023) for further details). No individuals were available for
 337 sampling from the 24°C treatment in the final month of the trial.

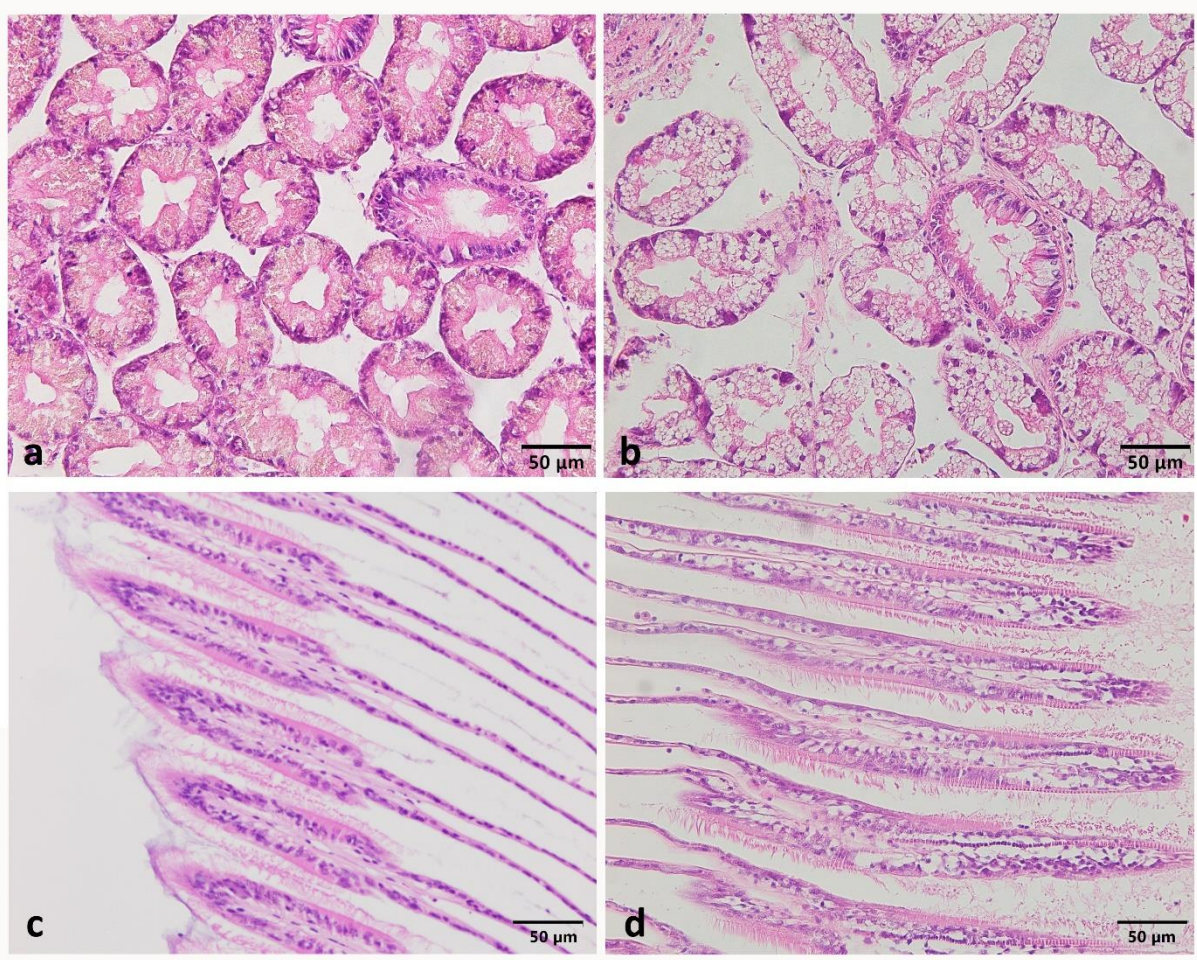
338 3.2. Tissue condition

339 3.2.1. Gill and digestive gland architecture

340 There were statistically significant histological differences in the appearance of the digestive
 341 gland tissue (Temp*Month p=0.008) and the gills (Temp*Month p=<0.001) in mussels among
 342 the three temperatures. There was an observable histological trend towards degradation in
 343 tissue architecture in the digestive gland and the gill condition of the mussels kept at all
 344 temperatures over time. These effects were only subtly different for the digestive gland there
 345 were individuals with 1) an increase in the intracellular spaces in the epithelium, which took
 346 on an increasingly thin and lacey appearance, 2) minor epithelial atrophy, and 3) epithelial
 347 sloughing in a few individuals, as well as a minimal increase in the amount of interstitial space

348 between the tubules as time progressed (Fig. 3a and b). For the gills, there were individuals
349 with some minor lifting between the chitin layer and the epithelium, increased cell separation,
350 minor hemocyte infiltration, in some cases loss of cilia and an accumulation of ceroid material
351 as time progressed (Fig. 3c and d).

352



353

354 Figure 3. General alterations to digestive gland tissues at trial start and trial end A) normal
355 digestive gland at trial initiation (graded as 0) B) altered digestive gland after 14 months at
356 24°C with expanded intracellular spaces (graded as 2). *Perna canaliculus* gill prior to
357 acclimation to the temperature treatments C) Healthy gill filaments (graded as 0) and D) Gill
358 filaments after 14 months with alterations to the gill architecture and no loss of cilia as
359 described in section 2.3.1.1. (graded as 2). Scale = 50µm.

360 3.2.2. Mantle and connective tissue

361 Semi-quantitative criteria of ceroid granules, focal hemocytosis, mantle storage cell coverage
362 and the extent of the sub-epithelial layer of the gut indicate that the differences between the

363 temperature treatments were linked to the month and were amplified over time (chronic time
 364 factor) (Table. 7).

365

366

367 Ceroid material was present in the majority of individuals and increased with both temperature
 368 and time (Temp*Month $p < 0.001$). Focal hemocytosis increased in intensity (, $p < 0.001$). with
 369 increasing temperature and was amplified over time ((Temp*Month $p = < 0.001$). There was a
 370 decrease in the coverage of the mantle storage cells which was also influenced by exposure
 371 time (Temp*Month $p < 0.001$). The subepithelial layer of the gut increased in thickness,
 372 extending into the digestive tubule tissue increasingly over time (Temp*Month $p < 0.001$)
 373 (Table. 7). In some individuals, the thickness was extensive and displaced the digestive tubules,
 374 resulting in the appearance of higher numbers, and increased density, of hemocytes.

375 Table 7: Heatmap of the semi-quantitative levels of ceroid material, focal hemocytosis, mantle
 376 storage cell coverage and the thickness of the subepithelial layer of the gut. Sampling month 3
 377 had fixation issues and, as such, only had an $n = 1$ at 17°C and $n = 3$ at 21°C ; no samples were
 378 available for the 24°C treatment. Sampling months 4 to 15 had $n = 8$ at 17°C , $n = 8$ at 21°C
 379 and $n = 8$ at 24°C for each sampling month (Table. 1). Where no data were collected, cells are
 380 blacked out.

381

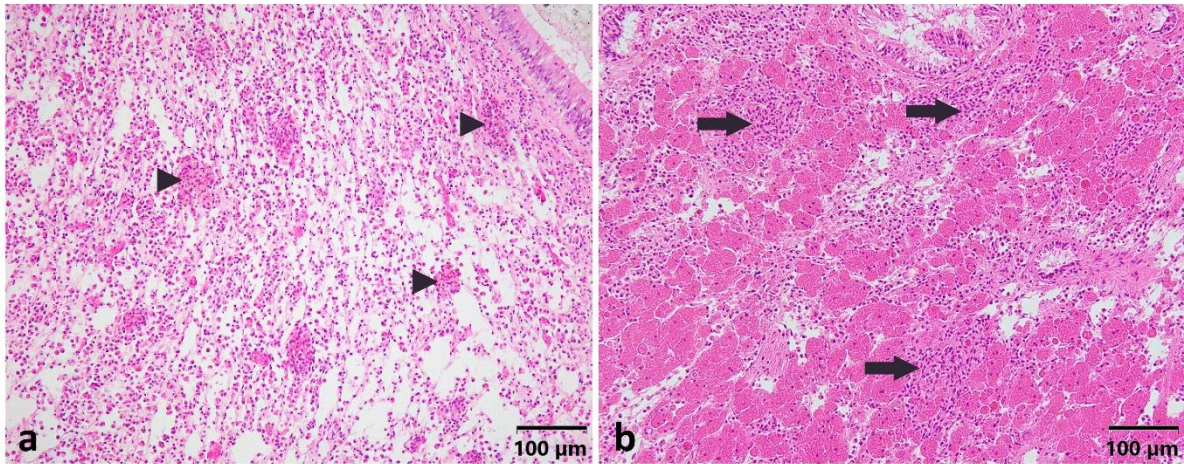


sub-epithelial gut layer

| | | | | | | | | | | | | | | | | |
|------|-----|-----|-----|-----|-----|-----|-----|-----|-----|-----|-----|-----|-----|-----|-----|-----|
| 17°C | 1.1 | 1.0 | 1.0 | 1.0 | 1.1 | 1.0 | 1.2 | 1.8 | 1.7 | 1.8 | 2.1 | 1.9 | 1.6 | 1.6 | 1.6 | 1.5 |
| 21°C | | 1.0 | 1.1 | 1.2 | 1.2 | 1.5 | 1.4 | 1.8 | 2.0 | 1.7 | 2.7 | 2.2 | 2.1 | 2.0 | 2.0 | 1.5 |
| 24°C | | | 1.1 | 1.1 | 1.2 | 1.3 | 1.1 | 1.8 | 2.0 | 2.3 | 2.4 | 2.4 | 2.3 | 2.4 | 2.1 | |

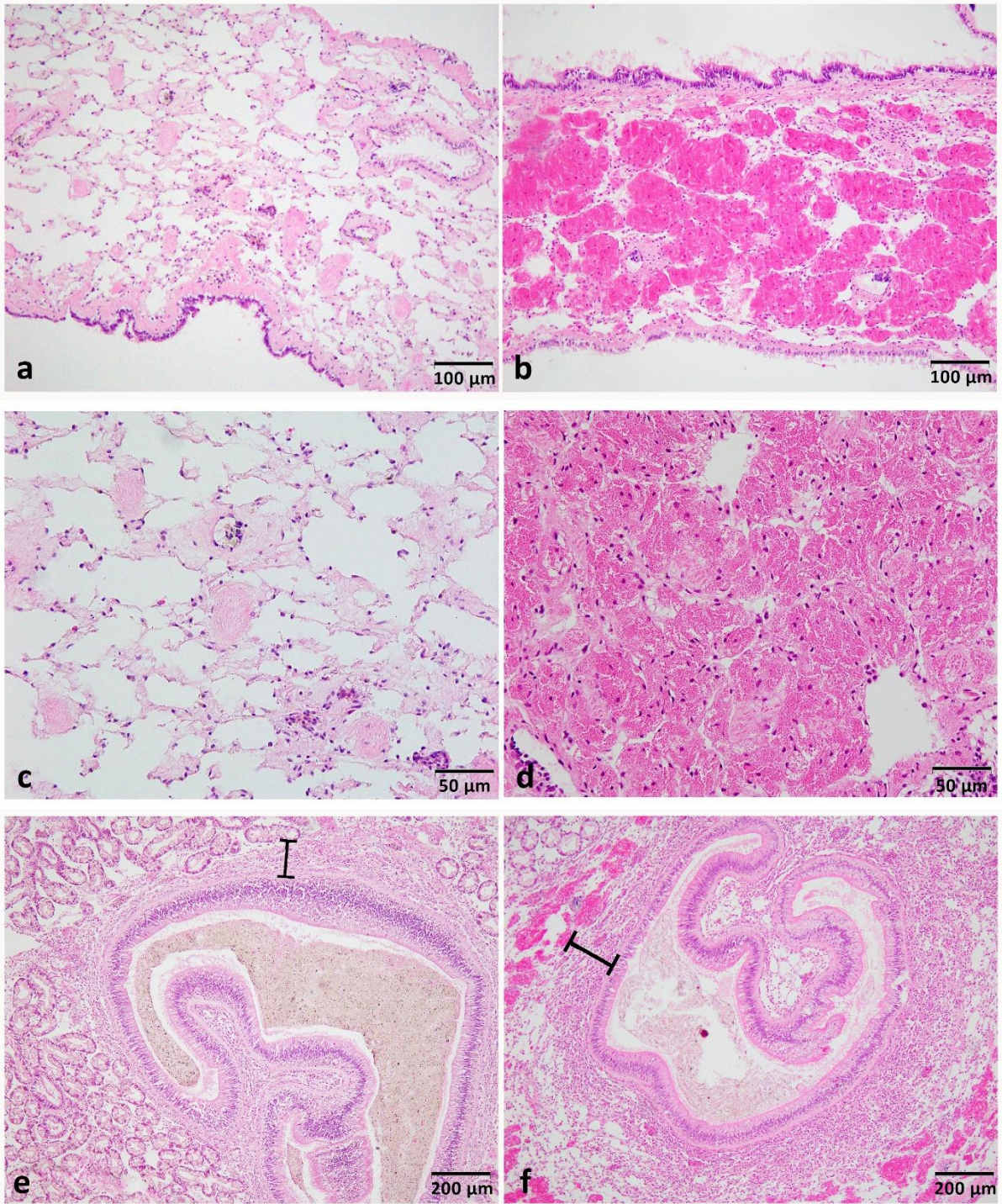
382

383



384

385 Figure 4. Ceroid level and the focal hemocytosis level 3. a) Severe level (score 3) of ceroid
386 material, multiple accumulations observed (arrowhead) b) Severe level (score 3) of focal
387 hemocytosis with multiple patches throughout the mantle tissue.



388

389 Figure 5. Mantle storage cell coverage and the subepithelial extent or thickness. a) Mantle
 390 tissue with a score of 0 no glycogen cells (granules) observed higher levels of hemocytes and
 391 connective tissue visible. b) Mantle tissue with a score of 2 with some patches of no storage
 392 cell tissue. c) Zoomed in mantle tissue of image a, only connective and some musculature
 393 tissue, with no visible glycogen granules. d) Zoomed in mantle tissue of image b showing
 394 densely packed storage cells e) Subepithelial layer of the intestine score of 1, low level of

395 hemocytes surrounding the epithelial layer of the intestine f) Subepithelial anomalous thick
396 layer (score of 3) around the intestine extending into the digestive tubule region, indicating
397 increased number of hemocytes surrounding the intestine Black line indicating subepithelial
398 layer

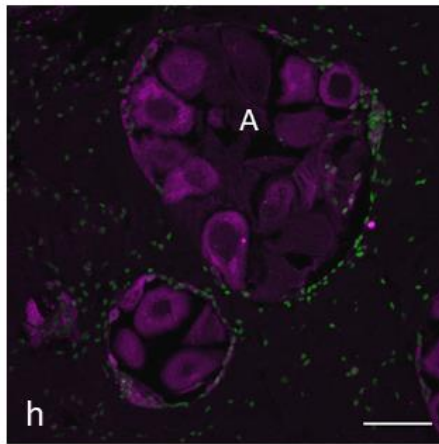
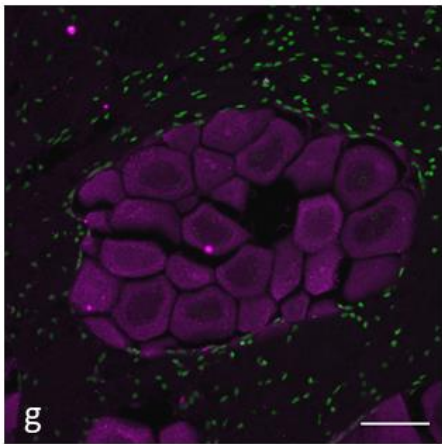
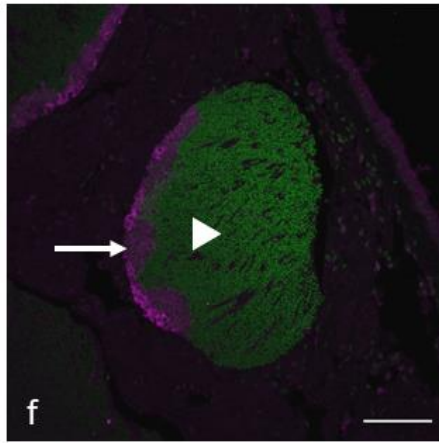
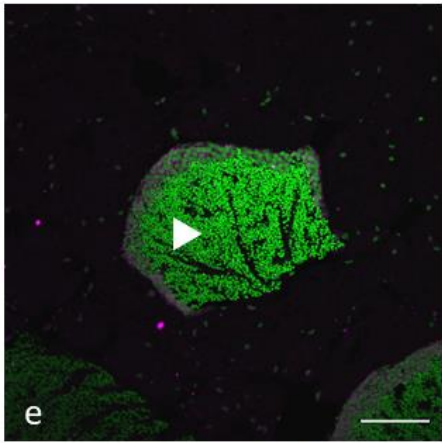
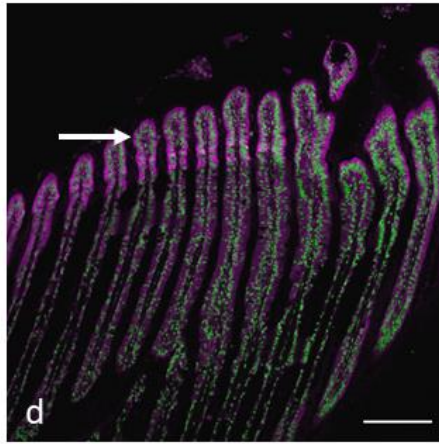
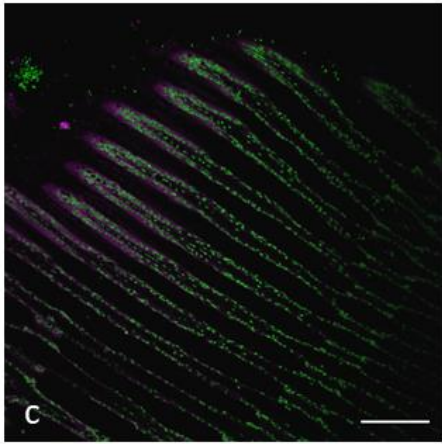
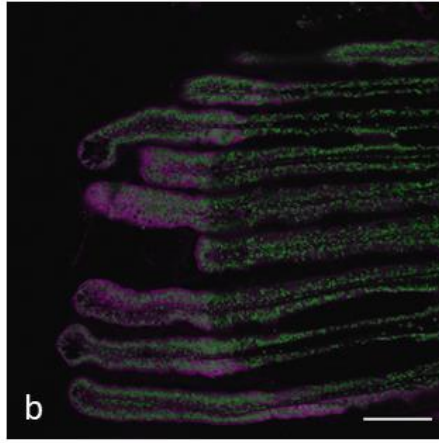
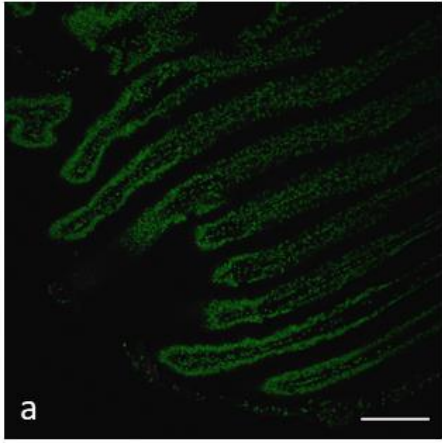
399

400 **3.2.3. Immunohistochemical exploration of HSP70 in gill and gonad: A preliminary**
401 **investigation.**

402 HSP70 presence in the gills appears to be localised to the epithelial cells at the frontal lateral
403 edge of each filament (Fig. 6 b, c, and d). The gill tissue suggested an increase in HSP70 protein
404 abundance with temperature in the ‘benign’ and elevated one-month post-temperature
405 acclimation. At this stage HSP70 proteins were not observed in the ambient ($17^{\circ}\text{C} \pm 2^{\circ}\text{C}$)
406 samples prior to the temperature increase for the experiment (Fig. 6 a).

407 HSP70 was suggested to be visible at elevated temperatures in both male and female gonads.
408 HSP70 in males appeared to be localised to spermatogonia and spermatocytes (Fig 6 e and f).
409 Female oocytes were highly abundant with HSP70 some oocytes displayed dense “patches”
410 with elevated temperatures (Fig. 6 g and h).

411



413 Figure 6. Preliminary investigations show the presence of HSP70 in gill and gonad tissue
414 determined by immunohistochemistry. a) gill tissue in 17°C temperature at month 2 prior to
415 the temperature elevation. Images b, c and d correspond to mussels that had been exposed for
416 1 month at 17, 21 and 24°C, respectively. Male reproductive follicle HSP70 (magenta) detected
417 to be associated to spermatogonia and spermatocytes, green is the mature sperm cell at 2
418 temperatures 1-month post-temperature increase e) 17°C and f) 21°C. Female reproductive
419 follicle at 2 temperatures 1-month post-temperature increase g) 17°C and h) 21°C with
420 increased atresic oocytes (A). Magenta (arrow) is the HSP70, and green (arrowhead) is the
421 nuclear material. Images A - D Scale bar= 100µm, Images E – H Scale bar= 50µm.

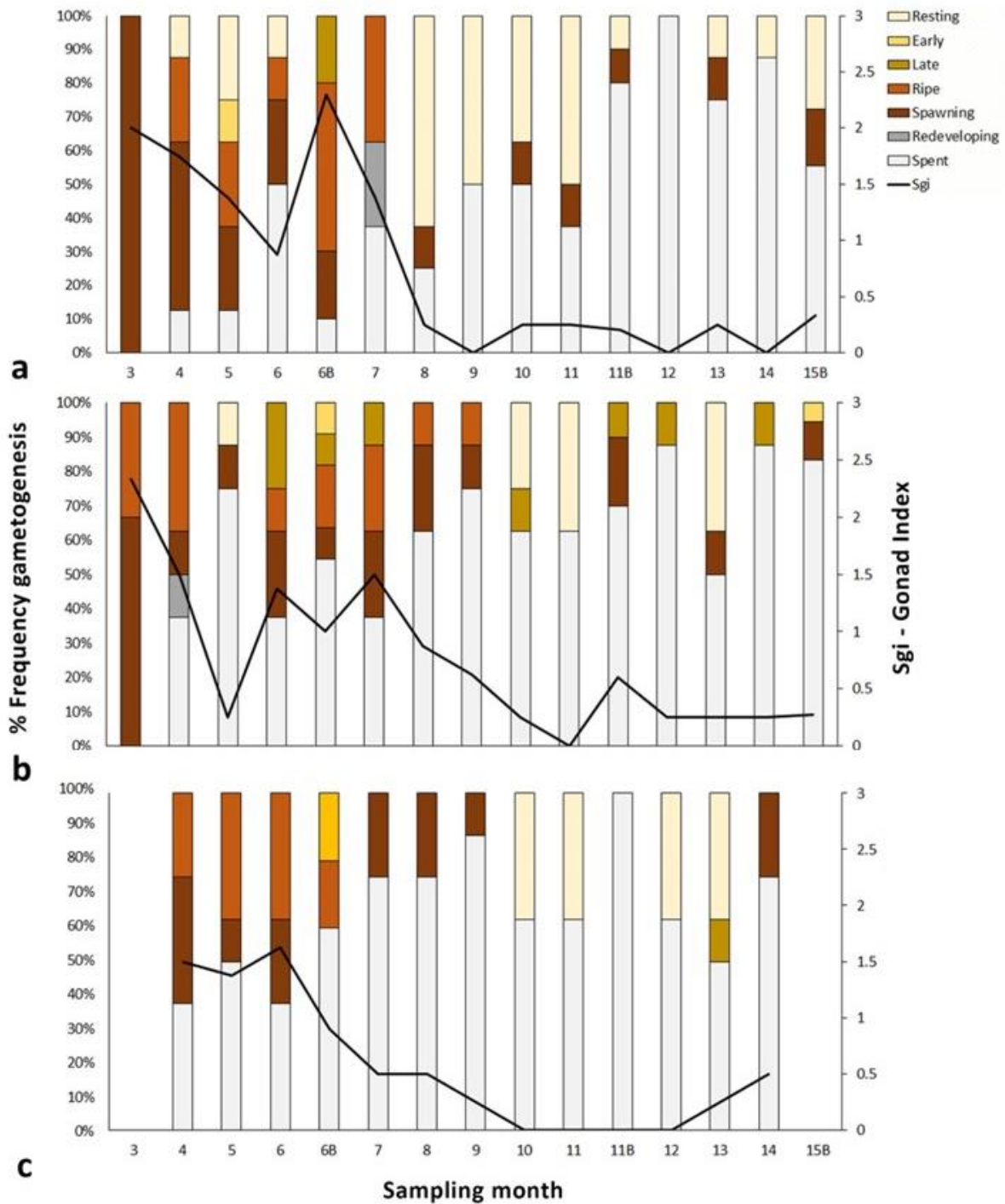
422

423 **3.3. Reproductive tissue**

424 Reproductive condition and gametogenesis appeared to have regressed over time, with much
425 of the population in either the spent or resting phase of gametogenesis by month eight. Initial
426 observational patterns suggest a natural decline (month 3 to 5 [21°C] and 6 [17°C]) from mature
427 individuals to the rest and spent phase, but no initiation of gametogenesis after the rest phase
428 (Fig. 8). There was no significant difference in gonadal development among the treatment
429 temperatures ($p=0.87$), nor in the interaction with month ($p=0.49$) (Fig. 7).

430

431

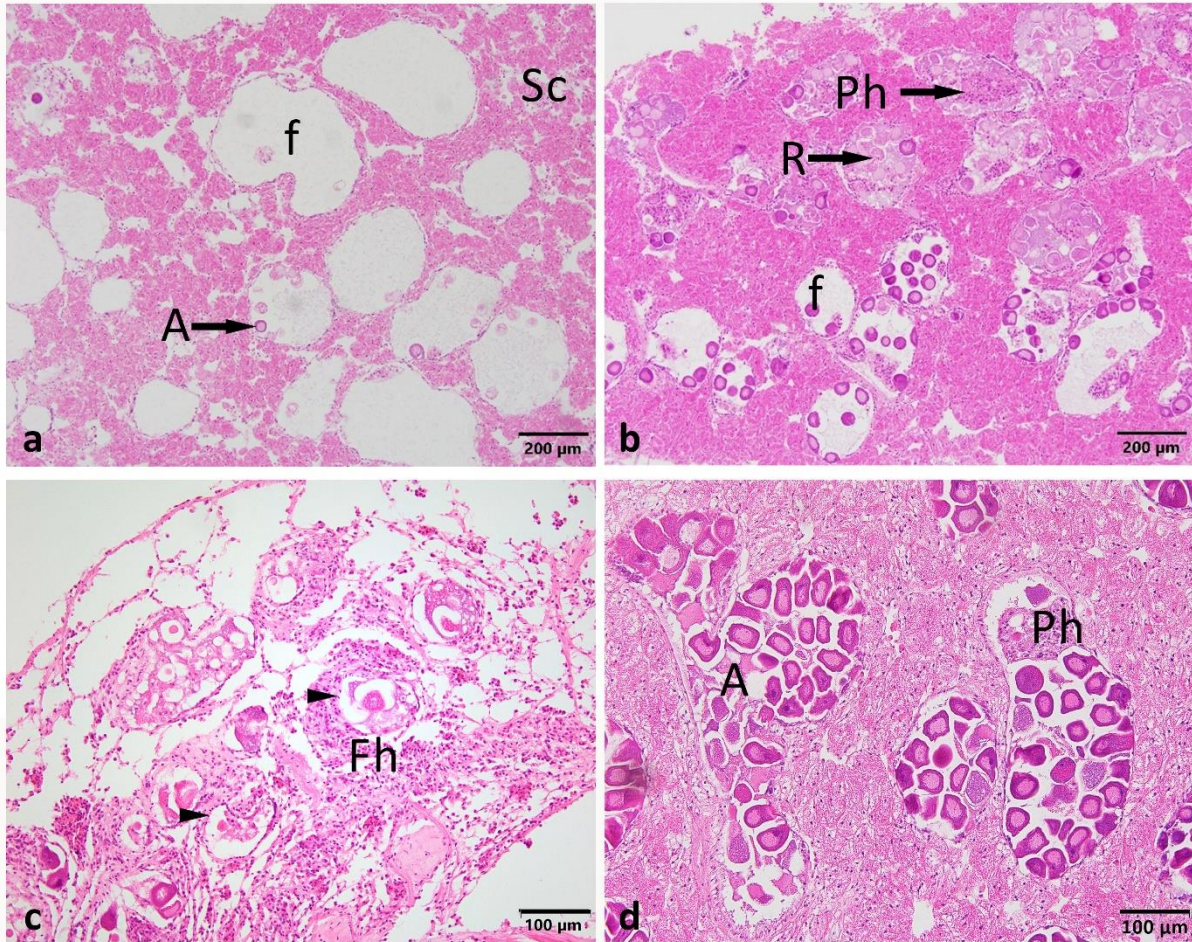


433

434 Figure 7. Gonad condition (proportion of sampled individuals in each development stage and
 435 overall gonad index) over 14 months exposure to a) 17°C, b) 21°C and c) 24°C. No data were
 436 collected at month 3 or 15B for 24°C.

437 There was no difference in the percent coverage of the follicles between the three temperatures
 438 ($p > 0.05$). Visual estimates of coverage correlated with objective image analysis quantification

439 with an R^2 value of 0.93. The female gonad tissue displayed various anomalies, particularly in
440 the spent category: 1) dilated follicles post-spawning, 2) dilated follicles with increased levels
441 of atresic debris, 3) increased numbers of follicles with atresic oocytes and some phagocytosis,
442 and 4) follicle regression and atrophy (Fig. 8).



443

444 Figure 8. Female reproductive tissue (mantle) showing: a) anomalous dilated post-spawning
445 follicle (f), b) dilated follicles with high levels of atresic debris, c) female mantle tissue with
446 low levels of storage cell (glycogen (Sc)), regressing and atrophying follicles (Arrow) and focal
447 hemocytosis (Fh) from a female in the 24°C treatment, d) full follicles with high level of atresic
448 oocytes (A), residual oocytes (R), and phagocytosis (Ph) within the follicle.

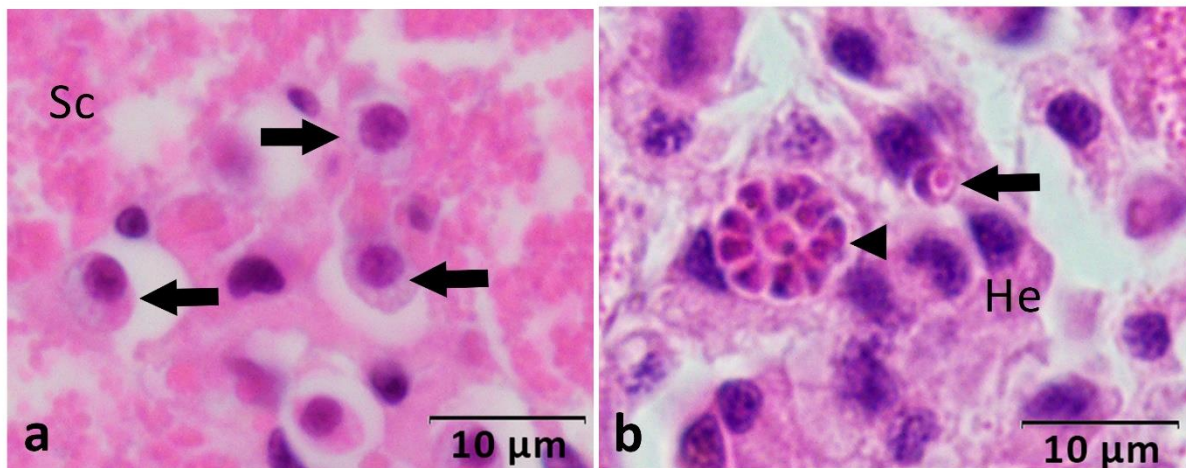
449 Due to the under-representation of females as a result of the majority of the sexes being
450 indeterminate across the three temperature treatments throughout the experiment, only
451 temperature was used as a factor, and female gonad images from each month were pooled. No
452 significant differences were observed among average proportion of immature oocytes ($F_{(3, 117)}=0.75$, $p=0.53$),
453 mature oocytes ($F_{(3, 117)}=0.85$, $p=0.47$), atresic oocytes ($F_{(3, 117)}=0.3$,

454 $p=0.82$), and atresic impact ($F_{(2, 109)}=0.4$, $p=0.76$). The average proportion of oocytes showing
 455 atresia, as well as overall atresic impact was observed to be high across the 17°C, 21°C and
 456 24°C temperatures ($62\% \pm 4.1$, $58\% \pm 4.3$, $60\% \pm 5.8$ and $74\% \pm 3.3$, $78\% \pm 2.9$, $75\% \pm 5.2$,
 457 respectively (mean percent \pm SD)).

458

459 3.4. Pathogens

460 There were several parasites detected in the *P. canaliculus* tissues. These included various
 461 unidentifiable ciliates and copepods, as well as a low prevalence of *Paravortex* and
 462 *Endozoicomonas* (Table. 8). The two most significant parasites observed were APX (Fig. 9a)
 463 and *Perkinsus olseni* (Fig. 9b).



464

465 Figure 9. *APX* and *Perkinsus olseni*. a) *APX* level 2: a group of 6 *APX* zoites within the mantle
 466 tissue no ceroid material or hemocytosis associated. b) *P. olseni* indicating level 3 displays
 467 different stages of the lifecycle trophozoite signet ring (arrow) and the rosette-like (arrowhead).
 468 *P.olseni* cells detectable across mantle and connective tissue of multiple tissues occasionally
 469 associated with hemocytosis.

470

471 Table 8. Population prevalence (%) of parasites found in the *P. canaliculus*. *APX* and *P. olseni*
 472 are excluded from this table and discuss in further detail below.

| | Temperature | | |
|------------------------------|-------------|------|------|
| | 17°C | 21°C | 24°C |
| <i>Endozoicomonas</i> | 0.0 | 1.6 | 1.9 |
| <i>Microsporidium rapuae</i> | 0.7 | 0.8 | 0.0 |
| <i>Paravortex</i> | 0.7 | 0.0 | 0.0 |

| | | | |
|----------|-----|-----|-----|
| Copepod | 2.1 | 1.6 | 2.9 |
| Ciliates | 1.4 | 0.0 | 0.0 |

473

474

475 APX was present in mussels in all three temperatures, and by month 4 of thermal exposure,
 476 APX approached 100% population prevalence. Small groups of APX were found scattered in
 477 the mantle, connective tissue of digestive glands and gut subepithelial as well as at the base of
 478 the gills and plicate organ near the kidney. Ceroid material was only associated occasionally at
 479 this level. No differences were identified among the sample time points, but a significant
 480 difference in the total population prevalence of APX was observed between mussels in the
 481 17°C and 24°C temperatures with prevalence being higher in the 24°C treatment ($t = 2.548$,
 482 $p=0.011$). *P. olseni* detection in mussels increased with temperature between 24°C compared
 483 to 17°C temperatures ($t = 6.051$, $p < 0.001$), and there was an interaction between temperature
 484 and time ($t = 4.058$, $p < 0.001$). The first detection of *P. olseni* in mussels occurred in 24°C, 5
 485 months after the start of thermal manipulation (i.e., month 7), which was 2 months earlier than
 486 detection in mussels at 21°C. *P. olseni* was detected in only 1 individual at 17°C towards the
 487 conclusion of the trial (Table 9). Focal hemocytosis (hemocytes - He) was associated to *P.*
 488 *olseni* cells.

489 Table 9. Population prevalence (%) of APX and *P. olseni* during 15 months of sampling at
 490 three temperatures 17°C, 21°C and 24°C (sample numbers available in Table 1)
 491

| | | Sampling month | | | | | | | | | | | | | | |
|------------------|------|----------------|-----|----|----|-----|-----|-----|-----|-----|-----|-----|-----|-----|-----|-----|
| | | 2b | 4 | 5 | 6 | 6b | 7 | 8 | 9 | 10 | 11 | 12b | 13 | 14 | 15 | 16b |
| <i>APX</i> | 17°C | 33 | 75 | 63 | 25 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 |
| | 21°C | | 100 | 75 | 50 | 80 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 88 | 100 |
| | 24°C | | 100 | 88 | 88 | 100 | 100 | 100 | 100 | 75 | 88 | 100 | 88 | 100 | 100 | |
| <i>P. olseni</i> | 17°C | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 13 | 0 | 0 |
| | 21°C | | 0 | 0 | 0 | 0 | 0 | 0 | 13 | 0 | 13 | 0 | 13 | 38 | 25 | 17 |
| | 24°C | | 0 | 0 | 0 | 0 | 38 | 25 | 25 | 50 | 63 | 50 | 25 | 50 | 25 | |

492

493 The APX intensity levels throughout this trial were quite low and, as such, only the first three
 494 semi-quantitative grades from Hine (2002) and Suong (2018) were required. The average
 495 intensities of APX across the three temperatures were 17°C = 1.3 ± 0.6 (n=109), 21°C = $1.4 \pm$
 496 0.5 (n=108) and 24°C = 1.1 ± 0.4 (n=90) (mean \pm SD). The average intensity of *P. olseni* was
 497 relatively low at 17°C = 1 (n=1), 21°C = 2.1 ± 0.9 (n=11) and 24°C = 1.8 ± 0.8 (n=28) (mean

498 ± SD). No lesions were identified in response to APX and minor hemocyte reactions were
499 present on occasion with *P. olseni* (Fig. 9).

500

501 **4. Discussion**

502 Prolonged exposure to constant temperature, even at the benign 17°C typical in summer was
503 associated with apparently deleterious histopathological alterations in digestive gland, gill and
504 gonad (The artificial constancy of laboratory trials and its potential confounding effect is
505 discussed in section 4.5). Higher temperatures and prolonged exposure were also associated
506 with increased prevalence of the pathogen *Perkinsus olseni*, suspension of gametogenesis, and
507 oocyte atresia (see below). Elevated temperatures, in addition, significantly impacted survival,
508 growth and immune response as also reported by (Ericson et al., 2023). The histopathological
509 alterations associated with elevated temperature may be valuable in assessing vulnerability to
510 environmental challenges such as those arising with climate change. Further refinement of
511 these potential indicators would contribute to improved health assessment of *P. canaliculus*
512 and the aquaculture industry it supports.

513 **4.1. Impacts on survival and growth.**

514 In a complementary study, Ericson et al. (2023) demonstrated that for *P. canaliculus*, prolonged
515 exposure to elevated temperatures (21°C and 24°C) was detrimental to both growth and
516 survival. This detrimental effect indicates that the optimal thermal range was exceeded with
517 consequential impacts on physiology, metabolism, immune response, and survival (Petes et al.,
518 2007; Dame and Kenneth, 2011; Delorme et al., 2021). In addition, there was a noticeable
519 difference in the development of byssus (anecdotal observation), where mussels at 24°C had
520 sparse and weak byssus development compared with mussels maintained at lower
521 temperatures. Although not part of the initial experiment this is worth noting and exploring
522 further in the future as mussels that prioritise growth, over byssal development in the natural
523 environment (i.e. rocky shores) risk dislodgement and increased mortality (Clarke, 1999;
524 Sebens et al., 2018; Roberts et al., 2021). The byssus threads were clipped four times
525 throughout the trial to weigh and measure mussels, thus, introducing a potential confounding
526 factor in terms of energy availability for growth, survival and immune response across all
527 temperature treatments. The energetic cost of byssal production for reattachment therefore
528 represented a compounding stressor for mussels within this study.

529 4.2. Tissue abnormalities

530 Chronic exposure to the three temperatures (17, 21 and 24°C) resulted in an accumulation of
531 ceroid (lipofuscin-like pigment) during the course of the trial. The ceroid accumulation in
532 response to thermal stress could be the result of an initial increase in metabolic rate, which
533 leads to increased oxygen consumption, followed by higher basal reactive oxygen species
534 (ROS) production, faster accumulation of peroxidised material and finally loss of cell function
535 (see Ericson et al., 2023). Therefore, the accumulation of ceroid indicates accelerated
536 senescence and potentially the advancement of the physiological age when compared with the
537 chronological age (Terman and Brunk, 1998; Basova et al., 2012). Typically, lipofuscin
538 production has been associated with aging and is also known to be a temperature dependent
539 process Miquel et al. (1976); (e.g. in clams Lomovasky et al., 2002), whereas ceroid is
540 considered to be a pigment associated to pathological conditions (Basova et al., 2012; Carella,
541 2015). There are several potential causes of ceroid accumulation, including immune response
542 and oxidative stress (Zarogian and Yevich, 1993). Increases in oxidative stress were observed
543 in mussels exposed to the same temperature treatments in the complementary study by Ericson
544 et al. (2023). Lipofuscin and ceroid are both end-products of lipid peroxidation and the
545 oxidation of proteins (Seehafer and Pearce, 2006; Basova et al., 2012; Carella, 2015), both
546 also have the ability to catalyse their own formation and drive further lipid peroxidation (Jung
547 et al. (2007). Due to the characteristic histological similarities between lipofuscin and ceroid
548 material, it is difficult to determine the difference between the two in this work. Therefore,
549 “ceroid” was used in this study to refer to the pigmented granules. Ceroid-lipofuscin
550 accumulation has been considered to be an outcome of inadequate intra-lysosomal digestion
551 from the phagocytosis process, these residual granules accumulate in the cell’s cytoplasm,
552 which can then be expelled into the connective tissue (Hendriks and Eestermans, 1986; Basova
553 et al., 2012; Hartenstein and Martinez, 2019). . In addition, accumulation may also indicate
554 previous and ongoing pathological conditions (i.e., APX and *Perkinsus sp.* exacerbating the
555 accumulation).

556 Elevated recruitment of hemocytes to the gut region at higher temperatures may indicate
557 defence against incoming pathogens or a response to altered gut microbiota and nutritional
558 affects. For example, increased aggregations of hemocytes around the gut have been observed
559 in *Crassostrea gigas* when exposed to the toxic algae *Alexandrium minutum* (Haberhorn et al.,
560 2010). While hemocytes can function as defence cells, they also fulfil digestion and nutrient
561 transport roles or support shell formation and tissue repair (e.g. Allam and Raftos, 2015; Rolton

562 and Ragg, 2020), the aggregation of hemocytes referred to as hemocytosis is a response limited
563 to tissue repair and pathogens (see section 4.3) (e.g. Haberkorn et al., 2010; Ben-Horin et al.,
564 2015). Across bivalve tissues, circulating hemocytes encapsulate and traffic pathogens such as
565 viral particles through the gut epithelium (Haberkorn et al., 2010; Ben-Horin et al., 2015) in a
566 process called diapedesis. This process has been suggested to be a defence response to protect
567 the tissues (e.g. Haberkorn et al. (2010). The digestive process, gut epithelial layer and mucosal
568 layers provide a strong defence against the physical environment and pathogens in healthy
569 individuals. However, during host stress, such as the thermal stress inflicted here, physiological
570 state and immunity may be compromised, allowing the entry of viruses, bacteria and parasites
571 through the gut tract and pallial organs, such as the gills. (Bower, 2006; Ben-Horin et al., 2015).

572 In addition to minor alterations to the gill architecture there also appeared to be expression
573 of the HSP70 protein with increasing temperature based on our preliminary observations.
574 HSP70 is a wide spectrum heat inducible protein in the HSP family which is expressed under
575 stress (Feder and Hofmann, 1999; Hu et al., 2022). Gills are directly exposed to the
576 environment and therefore detection of HSP70 is not unexpected as a “first layer of defence”
577 response to thermal stress. The increasing HSP70 expression with increasing temperature is
578 worth further research as production is associated with a high energetic cost (Feder and
579 Hofmann, 1999; Tomanek, 2010; Valenzuela-Castillo et al., 2019). Possibly though HSP70
580 production might have been beneficial for *P. canaliculus* exposed to higher temperatures,
581 which could explain the minimal gill architectural alterations observed. The inclusion of these
582 preliminary results is novel and promising for *P. canaliculus* and worth further investigation.

583 **4.3. Pathogen prevalence and influence on the host**

584 *Perkinsus olseni* was detected in mussels at month 7 at 24°C, month 9 at 21°C in the trial. The
585 intensity of the parasite, based on (Hine and Diggles, 2002) and (Kim et al., 2006) was
586 considered to be low and therefore potentially at the initial stage of infection. In association
587 with *P. olseni*, there was an increasing intensity of focal hemocytosis at the elevated
588 temperatures 21°C and 24°C. In addition, the presence of *P. olseni* was associated with
589 decreased energy reserves and immune responses such as increased intensity of focal ceroid
590 aggregations, focal hemocytosis, and encapsulation of parasites across different tissue types.
591 The parasite is known to have an affinity for waters above 20°C which commonly occur in the
592 North Island (Hine and Diggles, 2002) and the top of the South Island (Broekhuizen et al.,
593 2021). This temperature (20°C) may represent a threshold of natural occurrence of *P. olseni*
594 whereby infective cells are present in low concentrations in the environment, with temperatures

595 above this (e.g. 24°C from this research) accelerating initiation of infection (Lester, 1986;
596 Goggin and Lester, 1995). The present work corroborated this temperature effect.

597 Marine diseases are the result of complex host-parasite-environment interactions and there is
598 growing evidence to suggest that, in some cases, their increased prevalence and severity may
599 be associated with climate change (Harvell et al., 1999; Burge et al., 2014). Suboptimal
600 environmental conditions, such as temperature stress in combination with stressed hosts can
601 favour pathogen transmission and replication resulting in mortalities (Harvell et al., 1999).
602 *Perkinsus* sp. host interactions exemplify this. This protist endoparasite infects several species
603 worldwide including, Greenshell mussels, *Perna canaliculus* (Muznebin et al., 2022), the
604 clams *Austrovenus stutchburyi* (Dungan et al., 2007), *Paphies australis* (Ben-Horin et al.,
605 2015), abalone *Haliotis iris* (Hine and Diggles, 2002; Muznebin et al., 2021) and the oyster
606 *Crassostrea gigas* (Ben-Horin et al., 2015). Climate warming has been implicated in driving
607 increased spatial distribution of *Perkinsus* sp. and resulting disease events (Carella, 2015).

608 *Perkinsus* sp. inhibits phagocytosis and suppresses apoptosis of hemocytes in molluscs (OrdÁS
609 et al., 1999; Sunila and LaBanca, 2003; Hughes et al., 2010). Phagocytic inhibition allows the
610 parasites to infect and proliferate in circulating hemocytes, without being destroyed, and spread
611 to other tissues (e.g. OrdÁS et al., 1999; Sunila and LaBanca, 2003; Hughes et al., 2010; Ben-
612 Horin et al., 2015). Phagocytosis, is one of the main mechanisms for defence, it is a
613 temperature-dependent process (Oliver and Fisher, 1995) (Yu et al., 2009). The hemocytes in
614 the mussels at the warmer temperatures potentially tried to ingest the *P. olseni* cells initially,
615 without success. This process was then followed by parasite-mediated inhibition once infection
616 was established, with *Perkinsus* cells developing into replicative trophozoites. Partial
617 phagocytosing could have resulted in the inadequate intra-lysosomal digestion and therefore
618 further accumulation of ceroid material as described in the above section (Hendriks and
619 Eestermans, 1986; Hartenstein and Martinez, 2019).

620 As heatwaves progress, two outcomes are hypothesised, either pathogen intensity continues to
621 increase in response to favourable conditions, or the pathogen intensity diminishes allowing
622 for elimination in the host tissues during the cool winter months as previously suggested by
623 Goggin and Lester (1995). It remains a question of how *P. olseni* was acquired and whether
624 development to detectable levels was facilitated primarily by stress from the suboptimal
625 chronic conditions. In the current study, it was difficult to determine this or to establish the true
626 extent of the *P. olseni* infection, and whether *P. olseni* is detrimental to *P. canaliculus* and, if

627 so, at what intensity level. The occurrence of *P. olseni* was unexpected as it was not a pathogen-
628 focused trial and, as such, this research highlights the need for additional surveillance methods
629 and diagnostic tools including PCR and histology, when performing experiments. There is little
630 published information regarding *P. canaliculus* as a host for *P. olseni* (Muznebin et al., 2022)
631 and, considering that environmental fluctuations may be exacerbated by climate change, further
632 research is required on the disease progression and implications for the Greenshell mussel
633 industry and wild mussel beds. This should include work on farmed and wild populations to
634 reflect conditions that are experienced in the sea.

635 The decline in energy reserves (glycogen) within the mantle, and potentially the digestive
636 glands, is likely to explain the decline in *P. canaliculus* weight found by (Ericson et al., 2023)
637 in response to elevated temperature (21 and 24°C). Additionally, the thermal stress and
638 potentially the *P. olseni* infection, in combination, are likely to have resulted in the observed
639 decline in mantle energy reserves from month 10 onward at 24°C. However declines in weight-
640 based measurements, such as condition index have been observed in other mussels, such as
641 *Mytilus edulis* (Clements et al., 2018) and *P. canaliculus* (Venter et al., In Prep) primarily in
642 response to elevated temperatures. The energy available for growth, immune response and
643 reproduction is acquired through food assimilation and is stored as lipids and glycogen within
644 the mantle tissues, digestive glands, and muscle. These reserves are primarily used as energy
645 over the winter period, as well as for gametogenesis (Bayne et al., 1982; Hummel et al., 1989;
646 Fearman et al., 2009). Accumulation and use of reserve glycogen content is likely to vary with
647 environmental changes, such as extreme temperatures, pollution and starvation (Hummel et al.,
648 1989). Glycogen is the primary carbohydrate used for maintenance under stressful conditions
649 and has been used as an indicator of health (Bayne, 1976; Barber and Blake, 1981). The
650 depletion of glycogen in storage cells is of interest and possibly related to several factors,
651 including its allocation to processes, such as routine metabolic maintenance, shell and byssal
652 development, phagocytosis and immune responses, as well its appropriation by pathogens
653 (Cheng, 1983).

654 **4.4. Reproductive condition**

655 Histopathological investigations in this study have revealed that at initial sampling timepoints,
656 *P. canaliculus* mussels were passing the ripe and spawning stage and advancing towards the
657 post-spawning, spent stage. Prolonged experimental exposure, irrespective of temperature, was
658 observed to have impacts on reproductive condition in the present study. Gametogenesis was
659 suspended with increased atresia (oocyte degradation) in the majority of the mussels in all

660 temperature treatments, including the “benign” 17°C, which may result from a) the prolonged
661 temperature stress, b) without the quiescent winter phase, the initiation of gonadogenesis is
662 indistinguishable, regardless of the amount of food available, c) reproduction is the first
663 energetic process to be sacrificed when temperatures are outside the reproductive optimum and
664 d) an undetected mussel spawning occurred post-transfer into the system prior to trial initiation.

665 Successful reproduction is crucial for the survival of any species. It is an energy-demanding
666 process with its own physiological stressors. It has been found that sublethal conditions in the
667 environment can inhibit reproduction as many species will reallocate energy from gamete
668 production and move it to the somatic tissues for growth, defence and repair (Michalek-Wagner
669 and Willis, 2001; Petes et al., 2007). Heat shock proteins are one of the main defence and repair
670 mechanisms in organisms (Feder and Hofmann, 1999; Tomanek, 2010; Valenzuela-Castillo et
671 al., 2019). In this study, a seemingly higher HSP70 expression was observed in gonads of
672 mussels exposed to elevated temperatures, indicating that the mussels had activated specific
673 mechanisms to cope with the stress; as has been shown in other marine invertebrates (e.g. Nash
674 et al., 2019; Delorme et al., 2020a; Delorme et al., 2020b). Additionally, the HSP70 presence
675 in gonad tissue suggests that mussels responded to the heat stress by investing in loading the
676 gametes with HSP70 to protect their offspring against potential further increases in temperature
677 or other stressors. As a result of the promising preliminary observations further research and
678 targeted sampling is required to determine the mechanisms of parental investment in *P.*
679 *canaliculus* exposed to chronic heat stress.

680 The lack of noticeable spawning and increased oocyte atresia observed within this investigation
681 suggests a potential strategy of *P. canaliculus* to cope with prolonged stress and may indicate
682 a method of adaptation and survival. Atresia is the degeneration and reabsorption of oocytes
683 prior to a spawning event and is a strategy to recycle energy (Beninger, 2017; Chérel and
684 Beninger, 2017). Atresic oocytes are determined by their histological characterisations and
685 unusual fixation effects, such as shrunken, highly irregular shapes, cytoplasmic detachment
686 from membrane, lysing of the membrane and increased staining affinity (Beninger, 2017;
687 Chérel and Beninger, 2017). Atresia has been reported in a number of bivalve species, e.g. the
688 clam *Tapes philippinarum* (Chérel and Beninger, 2017), oyster *Crassostrea gigas* (Steele and
689 Mulcahy, 1999) and mussels *Mytilus edulis* (Pipe, 1987), *Mytilus galloprovincialis* (Ortiz-
690 Zarragoitia et al., 2011) and *Aulacomya atra* (Pérez et al., 2013). The probable causes of
691 increased incidence and intensities of atresia are adverse environmental conditions, including
692 pollution, starvation, and sub-optimal temperature (Galap et al., 1997; Steele and Mulcahy,

693 1999; Pérez et al., 2013; Vazquez et al., 2020). Atresia is a strategy to reallocate energy reserves
694 in order to cope with stress and energy depletion (Beninger, 2017). Observations of atresia in
695 this study are therefore expected, particularly when *P. canaliculus* were exposed to sub-optimal
696 high temperatures. Furthermore, the incidence of atresia prior to increasing the temperatures
697 was higher than expected and is of interest. Several questions remain for example, was the
698 incidence of atresia at this point representative of field conditions or a result of reallocation of
699 energy to byssal development following stripping from the farm ropes at initial collection time?
700 additionally was the atresia due to movement into the spent stage as a result of an undetected
701 spawning resulting in an additional stressor as well as a loss of energy. Failure to identify and
702 determine atresia within field and the laboratory populations will result in several implications
703 including over-estimations of fecundity and reproductive effort (Beninger, 2017).

704 Both reproduction and somatic growth reflect complex intrinsic and extrinsic interactions. The
705 reproductive cycle is controlled by both endogenous factors and environment, with temperature
706 being the driving force behind the initiation and the rate of gametogenesis (Michalek-Wagner
707 and Willis, 2001; Petes et al., 2007). Reproductive timing, synchronicity, fertilisation, and
708 recruitment success are influenced by temperature. Chronic stress can lead to a loss of
709 propagules, low gamete quality and spawning of pre-mature gametes that supply the adult
710 population (Walther et al., 2002; Philippart et al., 2003; Petes et al., 2007; Petes et al., 2008).
711 *P. canaliculus* is a dioecious broadcast spawning species with gonadal development and
712 gametogenesis occurring throughout the year (Alfaro et al., 2001; Buchanan, 2001). The
713 depletion of the glycogen content and the decline in growth (Ericson et al., 2023) in *P.*
714 *canaliculus* exposed to elevated temperatures during the present study suggests there is very
715 little residual energy after increased metabolism for both growth and reproduction, as observed
716 in other molluscan species (Fearman and Moltschaniwskyj, 2010). Present and future climate
717 change influences on environmental stressors are of major concern for *P. canaliculus* in the
718 aquaculture industry and in the wild. In terms of reproductive capability, it is crucial to
719 determine the true extent of oocyte atresia in *P. canaliculus* and how ongoing stressors, such
720 as marine heatwaves are likely to impact oocyte development further. Knowledge relating to
721 interactions between the reproductive condition and health of *P. canaliculus* is key for the
722 aquaculture industry, particularly as heatwaves increase in frequency and intensity (Beninger,
723 2017). Such expected changes to reproduction and spawning status of mussels could impact
724 the aquaculture industry by affecting not only their ability to plan harvesting operations, but
725 their ability to produce economically valuable products.

726 **4.5. Research implications**

727 It was surprising that the unchanging conditions within this study were ultimately detrimental
728 and 17°C, although considered as a benign temperature, does have negative effects over time.
729 This impact on overall health in the benign treatment was also observed in the immune stress
730 indicators in the complementary study by Ericson et al. (2023). It is yet to be determined
731 whether these effects on the benign temperature was due to captivity stressors, including
732 nutrition, density effects, lack of natural variation, and general holding conditions, although, a
733 similar effect of captivity has been considered and observed in other invertebrates such as
734 mytilids (Bayne and Thompson, 1970) and freshwater mussels (Roznere et al., 2021; Morin et
735 al., 2022). The thermal conditions in this study were maintained to eliminate treatment
736 variability and focus on temperature as a single stressor. Single stressor long-term studies such
737 as the work present herein are required to isolate the compounding effects of continuous stress
738 such as temperature elevations. While *P. canaliculus* mussels in their natural environment are
739 unlikely to be chronically exposed to elevated temperatures for the extended durations used in
740 this study the average sea surface temperatures are increasing. Although, predictions of climate
741 change are highly complex, impacts of increasing seawater temperatures on species phenology,
742 distribution and physiological performance are likely (Zippay and Helmuth, 2012; Shelmerdine
743 et al., 2017; Steeves et al., 2018). Therefore, the results of this work provide key knowledge
744 on health at a tissue level and caution is required when extrapolating laboratory data to field
745 conditions.

746 **4.6. Conclusions**

747 The present research identified changes in response to thermal stress and provides new insights
748 into the host-environment-pathogen interactome for *P. canaliculus* under changing
749 environments. Additionally, this is the first study that shows immunodetection of HSP70 in *P.*
750 *canaliculus*. The decline in growth, loss of byssal development and suspension of
751 gametogenesis in *P. canaliculus* suggests that, with increasing temperature and loss of seasonal
752 variability, there is insufficient energy available after metabolic maintenance and immune
753 defence in the laboratory conditions provided. In addition, the unexpected appearance of *P.*
754 *olseni* needs further investigation and could be indicative of potential range spread if average
755 temperatures and marine heatwave frequency continue to rise. Although, the HSP 70 technique
756 is not comprehensively studied within this research it is a novel technique for this species, the
757 small numbers of samples that were selected for IHC, do show promising protein expression

758 between temperature treatments in the gills and gonad and is worth future investigation.
759 Furthermore, the aetiological causes of the mortalities during the trial are also still to be
760 determined, with further investigations required to determine whether the mortalities were a
761 result of exhaustion of energy reserves, phagocytosis inhibition, increased toxicity in the tissues
762 from anaerobic metabolism, increased toxicity from the ceroid accumulation or other causes.
763 Timely sampling at point of death before decay had set in would have provided further
764 information for this investigation. This work highlights the potential effects of chronic thermal
765 stress if climate change and marine heatwave frequency continues to rise. Further work using
766 simultaneous laboratory and field studies incorporating realistic fluctuating and increased
767 temperatures, as well as baseline ambient sampling, will help to provide further beneficial
768 knowledge on the future impacts to *P. canaliculus* in response to climate change.

769

770 **Acknowledgements:**

771 We would like to give huge thanks to the technical staff at Cawthron Institute for the ongoing
772 maintenance and monitoring of the trial as well as the Aquaculture Biotechnology Research
773 Group at the Auckland University of technology for their support. We thank Lizenn Delisle
774 and Paula Casanovas for their ongoing support, and Rodney Roberts for his review and
775 comments. Lastly, we thank SPATnz for providing the Greenshell mussels used in this
776 experiment, Medlab Central Histopathology, Palmerston North, for their assistance with
777 histology processing and the Hugh Green Foundation for their ongoing support and funding of
778 the Hugh Green Cytometry Centre. This research was supported and funded by the New
779 Zealand Ministry for Business, Innovation and Employment, through the Cawthron Shellfish
780 Aquaculture Research Platform (Contract CAWX1801) and the Aquatic Animal Health
781 Programme (CAWX1707).

782

783 **5. References**

784 Alfaro, A.C., Jeffs, A.G., Hooker, S.H., 2001. Reproductive behavior of the green-lipped
785 mussel, *Perna canaliculus*, in Northern New Zealand. Bulletin of Marine Science 69, 1095-
786 1108
787 Alfaro, A.C., Young, T., 2018. Showcasing metabolomic applications in aquaculture: a
788 review. Reviews in Aquaculture 10, 135-152

789 Allam, B., Raftos, D., 2015. Immune responses to infectious diseases in bivalves. *J Invertebr*
790 *Pathol* 131, 121-136, 'doi'10.1016/j.jip.2015.05.005.

791 Angilletta Jr, M.J., Angilletta, M.J., 2009. Thermal adaptation: a theoretical and empirical
792 synthesis.

793 Barber, B.J., Blake, N.J., 1981. Energy storage and utilization in relation to gametogenesis in
794 *Argopecten irradians concentricus* (Say). *Journal of Experimental Marine Biology and*
795 *Ecology* 52, 121-134

796 Basova, L., Begum, S., Strahl, J., Sukhotin, A., Brey, T., Philipp, E., Abele, D., 2012. Age-
797 dependent patterns of antioxidants in *Arctica islandica* from six regionally separate
798 populations with different lifespans. *Aquatic Biology* 14, 141-152

799 Bayne, B., 1976. *Marine mussels, their ecology and physiology*. Cambridge University Press,
800 Cambridge

801 Bayne, B., Bubel, A., Gabbott, P., Livingstone, D., Lowe, D., Moore, M., 1982. Glycogen
802 utilisation and gametogenesis in *Mytilus edulis* L. *Marine Biology Letters* 3

803 Bayne, B.L., Thompson, R.J., 1970. Some physiological consequences of keeping *Mytilus*
804 *edulis* in the laboratory. *Helgoländer Wissenschaftliche Meeresuntersuchungen* 20, 526-552,
805 'doi'10.1007/bf01609927.

806 Behrens, E., Rickard, G., Rosier, S., Williams, J., Morgenstern, O., Stone, D., 2022.
807 Projections of Future Marine Heatwaves for the Oceans Around New Zealand Using New
808 Zealand's Earth System Model. *Frontiers in Climate* 4, 'doi'10.3389/fclim.2022.798287.

809 Ben-Horin, T., Bidegain, G., Huey, L., Narvaez, D.A., Bushek, D., 2015. Parasite
810 transmission through suspension feeding. *Journal of Invertebrate Pathology* 131, 155-176

811 Beninger, P.G., 2017. Caveat observator: the many faces of pre-spawning atresia in marine
812 bivalve reproductive cycles. *Marine Biology* 164, 163, 'doi'10.1007/s00227-017-3194-x.

813 Bignell, J.P., Dodge, M.J., Feist, S.W., Lyons, B., Martin, P.D., Taylor, N.G.H., Stone, D.,
814 Travalent, L., Stentiford, G.D., 2008. Mussel histopathology: effects of season, disease and
815 species. *Aquatic Biology* 2, 1-15, 'doi'10.3354/ab00031.

816 Bower, S.M., 2006. Synopsis of infectious diseases and parasites of commercially exploited
817 shellfish: Ciliates associated with Abalone. dfo,[https://www.dfo-mpo.gc.ca/science/aah-](https://www.dfo-mpo.gc.ca/science/aah-saa/diseases-maladies/ciliatesab-eng.html)
818 [saa/diseases-maladies/ciliatesab-eng.html](https://www.dfo-mpo.gc.ca/science/aah-saa/diseases-maladies/ciliatesab-eng.html).

819 Boyd, P.W., Lennartz, S.T., Glover, D.M., Doney, S.C., 2014. Biological ramifications of
820 climate-change-mediated oceanic multi-stressors. *Nature Climate Change* 5, 71-79,
821 'doi'10.1038/nclimate2441.

822 Broekhuizen, N., Plew, D.R., Pinkerton, M.H., Gall, M.G., 2021. Sea temperature rise over
823 the period 2002–2020 in Pelorus Sound, New Zealand – with possible implications for the
824 aquaculture industry. *New Zealand Journal of Marine and Freshwater Research* 55, 46-64,
825 'doi'10.1080/00288330.2020.1868539.

826 Buchanan, S., 2001. Measuring reproductive condition in the Greenshell™ mussel *Perna*
827 *canaliculus*. *New Zealand Journal of Marine and Freshwater Research* 35, 859-870,
828 'doi'10.1080/00288330.2001.9517048.

829 Burge, C.A., Mark Eakin, C., Friedman, C.S., Froelich, B., Hershberger, P.K., Hofmann,
830 E.E., Petes, L.E., Prager, K.C., Weil, E., Willis, B.L., Ford, S.E., Harvell, C.D., 2014.
831 Climate change influences on marine infectious diseases: implications for management and
832 society. *Ann Rev Mar Sci* 6, 249-277, 'doi'10.1146/annurev-marine-010213-135029.

833 Carella, F.F., S.W. Bignell, J.P. DeVico G. , 2015. Comparative pathology in bivalves:
834 Etiological agents and disease processes. *Journal of Invertebrate Pathology* 131, 13

835 Cheng, T.C., 1983. The role of lysosomes in molluscan inflammation. *American Zoologist*
836 23, 129-144

837 Chérel, D., Beninger, P.G., 2017. Oocyte Atresia Characteristics and Effect on Reproductive
838 Effort of Manila Clam *Tapes philippinarum* (Adams and Reeve, 1850). *Journal of Shellfish*
839 *Research* 36, 549-557, 'doi'10.2983/035.036.0302.

840 Chong, R.S.-M., 2022. Molluscan immunology, in: Kibenge, F.S.B., Baldisserotto, B.,
841 Chong, R.S.-M. (Eds.), *Aquaculture Pathophysiology*. Academic Press, pp. 383-
842 392,10.1016/b978-0-323-95434-1.00057-7.

843 Clarke, M., 1999. The effect of food availability on byssogenesis by the zebra mussel
844 (*Dreissena polymorpha Pallas*). *Journal of Molluscan Studies* 65, 327-333,
845 'doi'10.1093/mollus/65.3.327.

846 Clements, J.C., Hicks, C., Tremblay, R., Comeau, L.A., 2018. Elevated seawater temperature,
847 not pCO₂, negatively affects post-spawning adult mussels (*Mytilus edulis*) under food
848 limitation. *Conserv Physiol* 6, cox078, 'doi'10.1093/conphys/cox078.

849 Costa, P.M., 2018. Introduction, in: Costa, P.M. (Ed.), *The Handbook of Histopathological*
850 *Practices in Aquatic Environments*. Academic Press, pp. 1-20,10.1016/b978-0-12-812032-
851 3.00001-0.

852 Cule, E.S., Frankowski, M.D., 2021. ridge: Ridge Regression with Automatic Selection of the
853 Penalty Parameter. R package version 3.0, <https://CRAN.R-project.org/package=ridge>

854 Dame, R.F., Kenneth, M.J., 2011. *Ecology of marine bivalves: an ecosystem approach*.
855 Taylor & Francis,10.1201/b11220.

856 Dawber, C., Association, N.Z.M.F., Staff, N.Z.M.F.A., 2004. Lines in the water: A history of
857 greenshell mussel farming in New Zealand. River Press

858 de la Ballina, N.R., Maresca, F., Cao, A., Villalba, A., 2022. Bivalve Haemocyte
859 Subpopulations: A Review. *Frontiers in Immunology* 13, 826255,
860 'doi'10.3389/fimmu.2022.826255.

861 Delorme, N., Biessy, L., South, P., Zamora, L., Ragg, N., Burritt, D., 2020a. Stress-on-stress
862 responses of a marine mussel, *Perna canaliculus*: food limitation reduces the ability to cope
863 with heat stress in juveniles. *Marine Ecology Progress Series* 644, 105-117

864 Delorme, N.J., Frost, E.J., Sewell, M.A., 2020b. Effect of acclimation on thermal limits and
865 hsp70 gene expression of the New Zealand sea urchin *Evechinus chloroticus*. *Comp Biochem*
866 *Physiol A Mol Integr Physiol* 250, 110806, 'doi'10.1016/j.cbpa.2020.110806.

867 Delorme, N.J., Venter, L., Rolton, A., Ericson, J.A., 2021. Integrating animal health and
868 stress assessment tools using the green-lipped mussel *Perna canaliculus* as a case study.
869 *Journal of Shellfish Research* 40, 93-112

870 Dungan, C.F., Reece, K.S., Moss, J.A., Hamilton, R.M., Diggles, B.K., 2007. *Perkinsus*
871 *olseni* in vitro isolates from the New Zealand clam *Austrovenus stutchburyi*. *J Eukaryot*
872 *Microbiol* 54, 263-270, 'doi'10.1111/j.1550-7408.2007.00265.x.

873 Dunphy, B.J., Ragg, N.L., Collings, M.G., 2013. Latitudinal comparison of thermotolerance
874 and HSP70 production in F2 larvae of the greenshell mussel (*Perna canaliculus*). *Journal of*
875 *Experimental Biology* 216, 1202-1209, 'doi'10.1242/jeb.076729.

876 Dunphy, B.J., Ruggiero, K., Zamora, L.N., Ragg, N.L.C., 2018. Metabolomic analysis of
877 heat-hardening in adult green-lipped mussel (*Perna canaliculus*): A key role for succinic acid
878 and the GABAergic synapse pathway. *Journal of Thermal Biology* 74, 37-46,
879 'doi'<https://doi.org/10.1016/j.jtherbio.2018.03.006>.

880 Dunphy, B.J., Watts, E., Ragg, N.L.C., 2015. Identifying Thermally-Stressed Adult Green-
881 Lipped Mussels (*Perna canaliculus* Gmelin, 1791) via Metabolomic Profiling. *American*
882 *Malacological Bulletin* 33, 127-135, 129

883 Ericson, J.A., Venter, L., Copedo, J.S., Nguyen, V.T., Alfaro, A.C., Ragg, N.L.C., 2023.
884 Chronic heat stress as a predisposing factor in summer mortality of mussels, *Perna*
885 *canaliculus*. *Aquaculture* 564, 'doi'10.1016/j.aquaculture.2022.738986.

886 Fearman, J.-A., Bolch, C.J.S., Moltschaniwskyj, N.A., 2009. Energy storage and reproduction
887 in mussels, *Mytilus galloprovincialis*: The influence of diet quality. *Journal of Shellfish*
888 *Research* 28, 305-312, 308

889 Fearman, J., Moltschaniwskyj, N., 2010. Warmer temperatures reduce rates of gametogenesis
890 in temperate mussels, *Mytilus galloprovincialis*. *Aquaculture* 305, 20-25

891 Feder, M.E., Hofmann, G.E., 1999. Heat-shock proteins, molecular chaperones, and the stress
892 response: evolutionary and ecological physiology. *Annu Rev Physiol* 61, 243-282,
893 'doi'10.1146/annurev.physiol.61.1.243.

894 Filgueira, R., Guyondet, T., Comeau, L.A., Tremblay, R., 2016. Bivalve aquaculture-
895 environment interactions in the context of climate change. *Global Change Biology* 22, 3901-
896 3913, 'doi'10.1111/gcb.13346.

897 Fulda, S., Gorman, A.M., Hori, O., Samali, A., 2010. Cellular stress responses: cell survival
898 and cell death. *Int J Cell Biol* 2010, 214074, 'doi'10.1155/2010/214074.

899 Galap, C., Leboulenger, F., Grillot, J.P., 1997. Seasonal variations in biochemical
900 constituents during the reproductive cycle of the female dog cockle *Glycymeris glycymeris*.
901 *Marine Biology* 129, 625-634, 'doi'10.1007/s002270050205.

902 Galinou-Mitsoudi, S., Sinis, A.I., 1994. Reproductive Cycle and Fecundity of the Date
903 Mussel *Lithophaga Lithophaga* (Bivalvia: Mytilidae). *Journal of Molluscan Studies* 60, 371-
904 385, 'doi'10.1093/mollus/60.4.371.

905 Garrabou, J., Coma, R., Bensoussan, N., Bally, M., Chevaldonné, P., Cigliano, M., Diaz, D.,
906 Harmelin, J.G., Gambi, M.C., Kersting, D.K., Ledoux, J.B., Lejeusne, C., Linares, C.,
907 Marschal, C., PÉRez, T., Ribes, M., Romano, J.C., Serrano, E., Teixido, N., Torrents, O.,
908 Zabala, M., Zuberer, F., Cerrano, C., 2009. Mass mortality in Northwestern Mediterranean
909 rocky benthic communities: effects of the 2003 heat wave. *Global Change Biology* 15, 1090-
910 1103, 'doi'10.1111/j.1365-2486.2008.01823.x.

911 Goggin, C., Lester, R., 1995. *Perkinsus*, a protistan parasite of abalone in Australia: a review.
912 *Marine and Freshwater Research* 46, 639-646

913 Gosling, E., 2008. Bivalve molluscs: biology, ecology and culture. John Wiley & Sons

914 Haberkorn, H., Lambert, C., Le Goïc, N., Moal, J., Suquet, M., Guéguen, M., Sunila, I.,
915 Soudant, P., 2010. Effects of *Alexandrium minutum* exposure on nutrition-related processes
916 and reproductive output in oysters *Crassostrea gigas*. *Harmful Algae* 9, 427-439,
917 'doi'10.1016/j.hal.2010.01.003.

918 Hadfield, J.D., 2010. MCMC Methods for Multi-Response Generalized Linear Mixed
919 Models: TheMCMCglmmRPackage. *Journal of Statistical Software* 33, 1 - 22,
920 'doi'10.18637/jss.v033.i02.

921 Hartenstein, V., Martinez, P., 2019. Phagocytosis in cellular defense and nutrition: a food-
922 centered approach to the evolution of macrophages. *Cell Tissue Res* 377, 527-547,
923 'doi'10.1007/s00441-019-03096-6.

924 Harvell, C.D., Kim, K., Burkholder, J.M., Colwell, R.R., Epstein, P.R., Grimes, D.J.,
925 Hofmann, E.E., Lipp, E.K., Osterhaus, A.D., Overstreet, R.M., Porter, J.W., Smith, G.W.,
926 Vasta, G.R., 1999. Emerging marine diseases--climate links and anthropogenic factors.
927 *Science* 285, 1505-1510, 'doi'10.1126/science.285.5433.1505.

928 Hendriks, H.R., Eestermans, I.L., 1986. Phagocytosis and lipofuscin accumulation in lymph
929 node macrophages. *Mech Ageing Dev* 35, 161-167, 'doi'10.1016/0047-6374(86)90006-0.

930 Hine, M., Diggles, B., 2002. The distribution of *Perkinsus olseni* in New Zealand bivalve
931 molluscs. *Surveillance (Wellington)* 29, 8-11

932 Hine, P.M., 2002. Severe apicomplexan infection in the oyster *Ostrea chilensis*: a possible
933 predisposing factor in bonamiosis. *Dis Aquat Organ* 51, 49-60, 'doi'10.3354/dao051049.

934 Hobday, A.J., Alexander, L.V., Perkins, S.E., Smale, D.A., Straub, S.C., Oliver, E.C.J.,
935 Benthuyssen, J.A., Burrows, M.T., Donat, M.G., Feng, M., Holbrook, N.J., Moore, P.J.,
936 Scannell, H.A., Sen Gupta, A., Wernberg, T., 2016. A hierarchical approach to defining
937 marine heatwaves. *Progress in Oceanography* 141, 227-238,
938 'doi'10.1016/j.pocean.2015.12.014.

939 Hooper, C., Day, R., Slocombe, R., Benkendorff, K., Handlinger, J., Goulias, J., 2014.
940 Effects of severe heat stress on immune function, biochemistry and histopathology in farmed
941 Australian abalone (hybrid *Haliotis laevis* × *Haliotis rubra*). *Aquaculture* 432, 26-37,
942 'doi'10.1016/j.aquaculture.2014.03.032.

943 Houghton, J.E.T., Ding, Y., Griggs, D., Noguer, M., van der Linden, P., Dai, X., Maskell, M.,
944 Johnson, C., 2001. *Climate Change 2001: The Scientific Basis: Contribution of Working*
945 *Group I to the Third Assessment Report of the Intergovernmental Panel on Climate Change*
946 (IPCC). Cambridge University Press, p. 881

947 Howard, D.W., 2004. *Histological techniques for marine bivalve mollusks and crustaceans*.
948 NOAA, National Ocean Service, National Centers for Coastal Ocean Service, Center for
949 Coastal Environmental Health and Biomolecular Research,, Cooperative Oxford Laboratory
950 Hu, Z., Song, H., Feng, J., Zhou, C., Yang, M.-J., Shi, P., Yu, Z.-L., Li, Y.-R., Guo, Y.-J., Li,
951 H.-Z., Zhang, T., 2022. Massive Heat Shock Protein 70 Genes Expansion and Transcriptional
952 Signatures Uncover Hard Clam Adaptations to Heat and Hypoxia. *Frontiers in Marine*
953 *Science* 9, 'doi'10.3389/fmars.2022.898669.

954 Hughes, F.M., Foster, B., Grewal, S., Sokolova, I.M., 2010. Apoptosis as a host defense
955 mechanism in *Crassostrea virginica* and its modulation by *Perkinsus marinus*. *Fish &*
956 *Shellfish immunology* 29, 247-257

957 Hummel, H., de Wolf, L., Zurburg, W., Apon, L., Bogaards, R.H., van Ruitenburch, M., 1989.
958 The glycogen content in stressed marine bivalves: The initial absence of a decrease.
959 *Comparative Biochemistry and Physiology Part B: Comparative Biochemistry* 94, 729-733,
960 'doi'10.1016/0305-0491(89)90157-0.

961 Jeffs, A.G., Holland, R.C., Hooker, S.H., Hayden, B.J., 1999. Overview and Bibliography of
962 research on the Greenshell Mussel, *Perna canaliculus*, from New Zealand waters. *Journal of*
963 *Shellfish Research* 18, 347-360

964 Jung, T., Bader, N., Grune, T., 2007. Lipofuscin: formation, distribution, and metabolic
965 consequences. *Ann N Y Acad Sci* 1119, 97-111, 'doi'10.1196/annals.1404.008.

966 Kennedy, V.S., 2010. Reproduction in *Mytilus edulis aoteanus* and *Aulacomya maoriana*
967 (Mollusca: Bivalvia) from Taylors mistake, New Zealand. *New Zealand Journal of Marine*
968 *and Freshwater Research* 11, 255-267, 'doi'10.1080/00288330.1977.9515676.

969 Kim, Y., Ashton-Alcox, K.A., Powell, E.N., 2006. Histological techniques for marine bivalve
970 molluscs: update. NOAA/National Centers for Coastal Ocean Science

971 King, P.A., McGrath, D., Gosling, E.M., 2009. Reproduction and Settlement of *Mytilus*
972 *edulis* on an Exposed Rocky Shore in Galway Bay, West Coast of Ireland. *Journal of the*
973 *Marine Biological Association of the United Kingdom* 69, 355-365,
974 'doi'10.1017/s0025315400029465.

975 Knowles, G., Handler, J., Jones, B., Moltschaniwskyj, N., 2014. Hemolymph chemistry
976 and histopathological changes in Pacific oysters (*Crassostrea gigas*) in response to low
977 salinity stress. *J Invertebr Pathol* 121, 78-84, 'doi'10.1016/j.jip.2014.06.013.

978 Law, C.S., Rickard, G.J., Mikaloff-Fletcher, S.E., Pinkerton, M.H., Behrens, E., Chiswell,
979 S.M., Currie, K., 2017. Climate change projections for the surface ocean around New
980 Zealand. *New Zealand Journal of Marine and Freshwater Research* 52, 309-335,
981 'doi'10.1080/00288330.2017.1390772.

982 Lester, R., 1986. Abalone die-back caused by protozoan infection? *Australian Fisheries* 45,
983 26-27

984 Li, S., Alfaro, A.C., Nguyen, T.V., Young, T., Lulijwa, R., 2020. An integrated omics
985 approach to investigate summer mortality of New Zealand Greenshell™ mussels.
986 *Metabolomics* 16, 1-16

987 Lomovasky, B.J., Morriconi, E., Brey, T., Calvo, J., 2002. Individual age and connective
988 tissue lipofuscin in the hard clam *Eurhomalea exalbida*. Journal of Experimental Marine
989 Biology and Ecology 276, 83-94, 'doi'10.1016/s0022-0981(02)00240-x.

990 Lubchenco, J., Navarette, S., Tissot, B., Castilla, J.C., 1993. Lubchenco, J. L., S. A.
991 Navarette, B. N. Tissot, and J. C. Castilla. 1993. Possible ecological responses to global
992 climate change: nearshore benthic biota of northeastern Pacific coastal ecosystems. Chapter
993 12 In: H. A. Mooney, E. R. Fuentes, and B. I. Kronberg. eds. Earth System Responses to
994 Global Change: Contrasts between North and South America. Academic Press, San Diego,
995 Manduzio, H., Rocher, B., Durand, F., Galap, C., Leboulenger, F., 2005. The point about
996 oxidative stress in molluscs. Invertebrate Survival Journal 2, 91-104

997 Michalek-Wagner, K., Willis, B.L., 2001. Impacts of bleaching on the soft coral *Lobophytum*
998 *compactum* . I. Fecundity, fertilization and offspring viability. Coral Reefs 19, 231-239,
999 'doi'10.1007/s003380170003.

1000 Miquel, J., Lundgren, P.R., Bensch, K.G., Atlan, H., 1976. Effects of temperature on the life
1001 span, vitality and fine structure of *Drosophila melanogaster*. Mech Ageing Dev 5, 347-370,
1002 'doi'10.1016/0047-6374(76)90034-8.

1003 Mooney, H., Larigauderie, A., Cesario, M., Elmquist, T., Hoegh-Guldberg, O., Lavorel, S.,
1004 Mace, G.M., Palmer, M., Scholes, R., Yahara, T., 2009. Biodiversity, climate change, and
1005 ecosystem services. Current Opinion in Environmental Sustainability 1, 46-54,
1006 'doi'10.1016/j.cosust.2009.07.006.

1007 Morin, M., Jönsson, M., Wang, C.K., Craik, D.J., Degnan, S.M., Degnan, B.M., 2022.
1008 Crown-of-thorns starfish in captivity experience sustained large-scale changes in gene
1009 expression. bioRxiv, 2022.2007.2021.501052, 'doi'10.1101/2022.07.21.501052.

1010 Muznebin, F., Alfaro, A.C., Webb, S.C., 2021. Occurrence of *Perkinsus olseni* and other
1011 parasites in New Zealand black-footed abalone (*Haliotis iris*). New Zealand Journal of
1012 Marine and Freshwater Research 57, 261-281, 'doi'10.1080/00288330.2021.1984950.

1013 Muznebin, F., Alfaro, A.C., Webb, S.C., 2022. *Perkinsus olseni* and other parasites and
1014 abnormal tissue structures in New Zealand Greenshell™ mussels (*Perna canaliculus*) across
1015 different seasons. Aquaculture International 31, 547-582, 'doi'10.1007/s10499-022-00991-8.

1016 Nash, S., Johnstone, J., Rahman, M.S., 2019. Elevated temperature attenuates ovarian
1017 functions and induces apoptosis and oxidative stress in the American oyster, *Crassostrea*
1018 *virginica*: potential mechanisms and signaling pathways. Cell Stress Chaperones 24, 957-967,
1019 'doi'10.1007/s12192-019-01023-w.

1020 Nguyen, T.V., Alfaro, A.C., 2020. Metabolomics investigation of summer mortality in New
1021 Zealand Greenshell mussels (*Perna canaliculus*). *Fish Shellfish Immunol* 106, 783-791,
1022 'doi'10.1016/j.fsi.2020.08.022.

1023 Oliver, E.C.J., Benthuisen, J.A., Bindoff, N.L., Hobday, A.J., Holbrook, N.J., Mundy, C.N.,
1024 Perkins-Kirkpatrick, S.E., 2017. The unprecedented 2015/16 Tasman Sea marine heatwave.
1025 *Nat Commun* 8, 16101, 'doi'10.1038/ncomms16101.

1026 Oliver, L., Fisher, W., 1995. Comparative form and function of oyster *Crassostrea virginica*
1027 hemocytes from Chesapeake Bay (Virginia) and Apalachicola Bay (Florida). *Diseases of*
1028 *aquatic organisms* 22, 217-225

1029 OrdÁS, M.C., Novoa, B., Figueras, A., 1999. Phagocytosis inhibition of clam and mussel
1030 haemocytes by *Perkinsus atlanticus* secretion products. *Fish & Shellfish Immunology* 9, 491-
1031 503, 'doi'10.1006/fsim.1999.0208.

1032 Ortiz-Zarragoitia, M., Garmendia, L., Barbero, M.C., Serrano, T., Marigomez, I., Cajaraville,
1033 M.P., 2011. Effects of the fuel oil spilled by the Prestige tanker on reproduction parameters
1034 of wild mussel populations. *J Environ Monit* 13, 84-94, 'doi'10.1039/c0em00102c.

1035 Parmesan, C., Yohe, G., 2003. A globally coherent fingerprint of climate change impacts
1036 across natural systems. *Nature* 421, 37-42, 'doi'10.1038/nature01286.

1037 Pawley, J., 2006. *Handbook of biological confocal microscopy*. Springer Science & Business
1038 Media

1039 Perez-Cebrecos, M., Prieto, D., Blanco-Rayon, E., Izagirre, U., Ibarrola, I., 2022. Differential
1040 tissue development compromising the growth rate and physiological performances of mussel.
1041 *Mar Environ Res* 180, 105725, 'doi'10.1016/j.marenvres.2022.105725.

1042 Pérez, A.F., Boy, C.C., Curelovich, J.N., Pérez Barros, P., Calcagno, J.Á., 2013. Relationship
1043 between energy allocation and gametogenesis in *Aulacomya atra* (Bivalvia: Mytilidae) in a
1044 sub-Antarctic environment. 48, 459-469

1045 Petes, L.E., Menge, B.A., Murphy, G.D., 2007. Environmental stress decreases survival,
1046 growth, and reproduction in New Zealand mussels. *Journal of Experimental Marine Biology*
1047 *and Ecology* 351, 83-91

1048 Petes, L.E., Mouchka, M.E., Milston-Clements, R.H., Momoda, T.S., Menge, B.A., 2008.
1049 Effects of environmental stress on intertidal mussels and their sea star predators. *Oecologia*
1050 156, 671-680

1051 Philippart, C.J., van Aken, H.M., Beukema, J.J., Bos, O.G., Cadée, G.C., Dekker, R., 2003.
1052 Climate-related changes in recruitment of the bivalve *Macoma balthica*. *Limnology and*
1053 *Oceanography* 48, 2171-2185

1054 Pipe, R.K., 1987. Ultrastructural and cytochemical study on interactions between nutrient
1055 storage cells and gametogenesis in the mussel *Mytilus edulis*. *Marine Biology* 96, 519-528,
1056 'doi'10.1007/bf00397969.

1057 Portner, H.O., 2002. Climate variations and the physiological basis of temperature dependent
1058 biogeography: systemic to molecular hierarchy of thermal tolerance in animals. *Comp*
1059 *Biochem Physiol A Mol Integr Physiol* 132, 739-761, 'doi'10.1016/s1095-6433(02)00045-4.

1060 Quinn, G.P., Keough, M.J., 2012. *Experimental Design and Data Analysis for Biologists*.
1061 Cambridge University Press, Cambridge, 10.1017/cbo9780511806384.

1062 R Core Team, 2021. R: A language and environment for statistical computing. R Foundation
1063 for Statistical Computing. Vienna, Austria, <https://www.R-project.org/>

1064 Ren, J.S., Ragg, N.L.C., Cummings, V.J., Zhang, J., 2020. Ocean acidification and dynamic
1065 energy budget models: Parameterisation and simulations for the green-lipped mussel.
1066 *Ecological Modelling* 426, 109069, 'doi'10.1016/j.ecolmodel.2020.109069.

1067 Ren, J.S., Ross, A.H., 2005. Environmental influence on mussel growth: A dynamic energy
1068 budget model and its application to the greenshell mussel *Perna canaliculus*. *Ecological*
1069 *Modelling* 189, 347-362, 'doi'10.1016/j.ecolmodel.2005.04.005.

1070 Roberts, E.A., Newcomb, L.A., McCartha, M.M., Harrington, K.J., LaFramboise, S.A.,
1071 Carrington, E., Sebens, K.P., 2021. Resource allocation to a structural biomaterial: Induced
1072 production of byssal threads decreases growth of a marine mussel. *Functional Ecology* 35,
1073 1222-1239

1074 Rolton, A., Ragg, N.L.C., 2020. Green-lipped mussel (*Perna canaliculus*) hemocytes: A flow
1075 cytometric study of sampling effects, sub-populations and immune-related functions. *Fish*
1076 *Shellfish Immunol* 103, 181-189, 'doi'10.1016/j.fsi.2020.05.019.

1077 Roznere, I., Sinn, B.T., Daly, M., Watters, G.T., 2021. Freshwater mussels (Unionidae)
1078 brought into captivity exhibit up-regulation of genes involved in stress and energy
1079 metabolism. *Sci Rep* 11, 2241, 'doi'10.1038/s41598-021-81856-7.

1080 RStudio Team, 2021. RStudio: Integrated Development Environment for R, RStudio, PBC

1081 Rubio-Portillo, E., Izquierdo-Munoz, A., Gago, J.F., Rossello-Mora, R., Anton, J., Ramos-
1082 Espla, A.A., 2016. Effects of the 2015 heat wave on benthic invertebrates in the Tabarca
1083 Marine Protected Area (southeast Spain). *Mar Environ Res* 122, 135-142,
1084 'doi'10.1016/j.marenvres.2016.10.004.

1085 Salinger, M.J., Diamond, H.J., Behrens, E., Fernandez, D., Fitzharris, B.B., Herold, N.,
1086 Johnstone, P., Kerckhoffs, H., Mullan, A.B., Parker, A.K., 2020a. Unparalleled coupled

1087 ocean-atmosphere summer heatwaves in the New Zealand region: drivers, mechanisms and
1088 impacts. *Climatic Change* 162, 485-506

1089 Salinger, M.J., Diamond, H.J., Behrens, E., Fernandez, D., Fitzharris, B.B., Herold, N.,
1090 Johnstone, P., Kerckhoffs, H., Mullan, A.B., Parker, A.K., Renwick, J., Scofield, C., Siano,
1091 A., Smith, R.O., South, P.M., Sutton, P.J., Teixeira, E., Thomsen, M.S., Trought, M.C.T.,
1092 2020b. Unparalleled coupled ocean-atmosphere summer heatwaves in the New Zealand
1093 region: drivers, mechanisms and impacts. *Climatic Change* 162, 485-506,
1094 'doi'10.1007/s10584-020-02730-5.

1095 Salinger, M.J., Renwick, J., Behrens, E., Mullan, A.B., Diamond, H.J., Sirguey, P., Smith,
1096 R.O., Trought, M.C.T., Alexander, L., Cullen, N.J., Fitzharris, B.B., Hepburn, C.D., Parker,
1097 A.K., Sutton, P.J., 2019. The unprecedented coupled ocean-atmosphere summer heatwave in
1098 the New Zealand region 2017/18: drivers, mechanisms and impacts. *Environmental Research*
1099 *Letters* 14, 044023, 'doi'10.1088/1748-9326/ab012a.

1100 Schindelin, J., Arganda-Carreras, I., Frise, E., Kaynig, V., Longair, M., Pietzsch, T.,
1101 Preibisch, S., Rueden, C., Saalfeld, S., Schmid, B., Tinevez, J.Y., White, D.J., Hartenstein,
1102 V., Eliceiri, K., Tomancak, P., Cardona, A., 2012. Fiji: an open-source platform for
1103 biological-image analysis. *Nat Methods* 9, 676-682, 'doi'10.1038/nmeth.2019.

1104 Sebens, K.P., Sara, G., Carrington, E., 2018. Estimation of fitness from energetics and life-
1105 history data: An example using mussels. *Ecol Evol* 8, 5279-5290, 'doi'10.1002/ece3.4004.

1106 Seehafer, S.S., Pearce, D.A., 2006. You say lipofuscin, we say ceroid: defining
1107 autofluorescent storage material. *Neurobiol Aging* 27, 576-588,
1108 'doi'10.1016/j.neurobiolaging.2005.12.006.

1109 Seuront, L., Nicastro, K.R., Zardi, G.I., Goberville, E., 2019. Decreased thermal tolerance
1110 under recurrent heat stress conditions explains summer mass mortality of the blue mussel
1111 *Mytilus edulis*. *Sci Rep* 9, 17498, 'doi'10.1038/s41598-019-53580-w.

1112 Shelmerdine, R.L., Mouat, B., Shucksmith, R.J., 2017. The most northerly record of feral
1113 Pacific oyster *Crassostrea gigas* (Thunberg, 1793) in the British Isles. *BioInvasions Records*
1114 6, 57-60

1115 Smale, D.A., Wernberg, T., Oliver, E.C.J., Thomsen, M., Harvey, B.P., Straub, S.C.,
1116 Burrows, M.T., Alexander, L.V., Benthuyssen, J.A., Donat, M.G., Feng, M., Hobday, A.J.,
1117 Holbrook, N.J., Perkins-Kirkpatrick, S.E., Scannell, H.A., Sen Gupta, A., Payne, B.L.,
1118 Moore, P.J., 2019. Marine heatwaves threaten global biodiversity and the provision of
1119 ecosystem services. *Nature Climate Change* 9, 306-312, 'doi'10.1038/s41558-019-0412-1.

1120 Steele, S., Mulcahy, M., 1999. Gametogenesis of the oyster *Crassostrea gigas* in southern
1121 Ireland. *Journal of the Marine Biological Association of the United Kingdom* 79, 673-686
1122 Steeves, L., Filgueira, R., Guyondet, T., Chassé, J., Comeau, L., 2018. Past, Present, and
1123 Future: Performance of Two Bivalve Species Under Changing Environmental Conditions.
1124 *Frontiers in Marine Science* 5, 184, 'doi'10.3389/fmars.2018.00184.
1125 Stenton-Dozey, J.M.E., Heath, P., Ren, J.S., Zamora, L.N., 2020. New Zealand aquaculture
1126 industry: research, opportunities and constraints for integrative multitrophic farming. *New*
1127 *Zealand Journal of Marine and Freshwater Research* 55, 265-285,
1128 'doi'10.1080/00288330.2020.1752266.
1129 Sunila, I., LaBanca, J., 2003. Apoptosis in the pathogenesis of infectious diseases of the
1130 eastern oyster *Crassostrea virginica*. *Diseases of Aquatic Organisms* 56, 163-170,
1131 'doi'10.3354/dao056163.
1132 Suong, N.T., Banks, J.C., Webb, S.C., Jeffs, A., Wakeman, K.C., Fidler, A., 2018. PCR test
1133 to specifically detect the apicomplexan 'X' (APX) parasite found in flat oysters *Ostrea*
1134 *chilensis* in New Zealand. *Dis Aquat Organ* 129, 199-205, 'doi'10.3354/dao03244.
1135 Sutton, P.J.H., Bowen, M., 2019. Ocean temperature change around New Zealand over the
1136 last 36 years. *New Zealand Journal of Marine and Freshwater Research* 53, 305-326,
1137 'doi'10.1080/00288330.2018.1562945.
1138 Terman, A., Brunk, U.T., 1998. Lipofuscin: mechanisms of formation and increase with age.
1139 *APMIS* 106, 265-276, 'doi'10.1111/j.1699-0463.1998.tb01346.x.
1140 Thomas, C.D., Cameron, A., Green, R.E., Bakkenes, M., Beaumont, L.J., Collingham, Y.C.,
1141 Erasmus, B.F., De Siqueira, M.F., Grainger, A., Hannah, L., Hughes, L., Huntley, B., Van
1142 Jaarsveld, A.S., Midgley, G.F., Miles, L., Ortega-Huerta, M.A., Peterson, A.T., Phillips, O.L.,
1143 Williams, S.E., 2004. Extinction risk from climate change. *Nature* 427, 145-148,
1144 'doi'10.1038/nature02121.
1145 Thomsen, M.S., Mondardini, L., Alestra, T., Gerrity, S., Tait, L., South, P.M., Lilley, S.A.,
1146 Schiel, D.R., 2019. Local Extinction of Bull Kelp (*Durvillaea* spp.) Due to a Marine
1147 Heatwave. *Frontiers in Marine Science* 6, 'doi'10.3389/fmars.2019.00084.
1148 Tomanek, L., 2010. Variation in the heat shock response and its implication for predicting the
1149 effect of global climate change on species' biogeographical distribution ranges and metabolic
1150 costs. *J Exp Biol* 213, 971-979, 'doi'10.1242/jeb.038034.
1151 Trump, B.F., Berezesky, I.K., Chang, S.H., Phelps, P.C., 1997. The pathways of cell death:
1152 oncosis, apoptosis, and necrosis. *Toxicol Pathol* 25, 82-88,
1153 'doi'10.1177/019262339702500116.

1154 Tuckett, C.A., de Bettignies, T., Fromont, J., Wernberg, T., 2017. Expansion of corals on
1155 temperate reefs: direct and indirect effects of marine heatwaves. *Coral Reefs* 36, 947-956,
1156 'doi'10.1007/s00338-017-1586-5.

1157 Valenzuela-Castillo, A., Sánchez-Paz, A., Castro-Longoria, R., López-Torres, M.A.,
1158 Grijalva-Chon, J.M., 2019. Hsp70 function and polymorphism, its implications for mollusk
1159 aquaculture: a review. *Latin american journal of aquatic research* 47, 224-231

1160 Vazquez, N., Frizzera, A., Cremonte, F., 2020. Diseases and parasites of wild and cultivated
1161 mussels along the Patagonian coast of Argentina, southwest Atlantic Ocean. *Dis Aquat Organ*
1162 139, 139-152, 'doi'10.3354/dao03467.

1163 Venter, L., Alfaro, A., C, Ragg, N.L., C, Delorme, N.J., Ericson, J.A., In Prep. The effect of
1164 simulated marine heatwaves on green-lipped mussels, *Perna canaliculus*: A near-natural
1165 experimental approach.

1166 Walther, G.R., Post, E., Convey, P., Menzel, A., Parmesan, C., Beebee, T.J., Fromentin, J.M.,
1167 Hoegh-Guldberg, O., Bairlein, F., 2002. Ecological responses to recent climate change.
1168 *Nature* 416, 389-395, 'doi'10.1038/416389a.

1169 Yu, J.H., Song, J.H., Choi, M.C., Park, S.W., 2009. Effects of water temperature change on
1170 immune function in surf clams, *Macra veneriformis* (Bivalvia: Mactridae). *J Invertebr Pathol*
1171 102, 30-35, 'doi'10.1016/j.jip.2009.06.002.

1172 Zannella, C., Mosca, F., Mariani, F., Franci, G., Folliero, V., Galdiero, M., Tiscar, P.G.,
1173 Galdiero, M., 2017. Microbial Diseases of Bivalve Mollusks: Infections, Immunology and
1174 Antimicrobial Defense. *Mar Drugs* 15, 182, 'doi'10.3390/md15060182.

1175 Zaroogian, G., Yevich, P., 1993. Cytology and biochemistry of brown cells in *Crassostrea*
1176 *virginica* collected at clean and contaminated stations. *Environmental Pollution* 79, 191-197

1177 Zippay, M.L., Helmuth, B., 2012. Effects of temperature change on mussel, *Mytilus*. *Integr*
1178 *Zool* 7, 312-327, 'doi'10.1111/j.1749-4877.2012.00310.x.

1179

1180