

1 [Effects of sediment disturbance by the heart urchin \*Echinocardium cordatum\*](#)  
2 [on the sediment–seawater solute exchange: an exclusion experiment](#)

3 Roen McLeod, Michelle Simone and Kay Vopel\*

4 School of Science, Auckland University of Technology, Auckland, New Zealand

5 \*Correspondence: Kay Vopel, [kay.vopel@aut.ac.nz](mailto:kay.vopel@aut.ac.nz)

6 Keywords: Bioturbation, sediment–water solute exchange, oxygen, nitrogen

7 [Abstract](#)

8 Spatangoid heart urchins are dominant bioturbators in marine soft-sediment ecosystems  
9 worldwide. Their repeated sediment reworking prevents biogeochemical sediment  
10 stratification and colonisation by other species, with implications for sedimentary reaction  
11 processes that affect the local sediment–seawater solute exchange. Here, we used a simple  
12 exclusion experiment to investigate how a subtidal *Echinocardium cordatum* population  
13 ( $18.2 \pm 6.7$  individuals  $m^{-2}$ ), foraging at an individual speed of  $\sim 45$  cm per day affects the  
14 sediment–seawater solute exchange. To do so, we removed all heart urchins from eight one-  
15 meter-diameter areas of the 10-m deep seafloor of Man O’War Bay, Hauraki Gulf, New  
16 Zealand, and prevented recolonisation and hence sediment reworking for 56 days.  
17 Subsequently, we measured the sediment–seawater exchange of  $O_2$ ,  $NO_3^-$ ,  $NO_2^-$ ,  $NH_4^+$ , and  
18  $N_2$  both within and outside the exclusion areas, under light or dark conditions, and found no  
19 difference. The absence of a legacy effect of foraging *E. cordatum* after their removal  
20 suggests that, at least in this habitat, their influence on the sediment–seawater solute  
21 exchange may be limited to sediment being displaced in the immediate surrounding of the  
22 urchin. This unexpected result underlines the importance of evaluating the influence of  
23 bioturbators on the sediment–seawater solute exchange in the context of local  
24 environmental conditions, animal behaviour, and population characteristics.

25

## 26 Introduction

27 Macrofaunal bioturbation profoundly influences the mineralisation of carbon in seafloor  
28 sediment by displacing particles and solutes (Aller and Aller, 1998; Banta et al., 1999; Solan  
29 and Wingham, 2005; Kristensen et al., 2012). Persistent sediment disturbances by, for  
30 example, bottom trawling or deposition of organic fish farm waste, can alter the composition  
31 of the resident macrofaunal assemblage and therefore affect carbon remineralisation rates  
32 and the associated sediment–seawater exchange of dissolved carbon and nitrogen. Because  
33 this exchange links seafloor with pelagic ecosystem functions, changes in sediment  
34 bioturbation can have far-reaching consequences for coastal primary and secondary  
35 production (Forster et al., 1995; Lohrer et al., 2004; Volkenborn et al., 2007).

36 Heart urchins of the genus *Echinocardium* Gray, 1825 have a disproportionately large  
37 influence on sedimentary transport and reaction processes (Osinga et al., 1997; Lohrer et al.,  
38 2004, 2005, 2015). They are globally widespread through soft sediment environments,  
39 inhabiting intertidal mud or sand flats to abyssal plains, and can form up to 60% of the  
40 macrobenthic biomass in some ecosystems (Nakamura, 2001; De Ridder & Saucède, 2020).  
41 Functioning as key bioturbators, they alter the physical and biological structure of their  
42 surrounding sediment both directly, by mixing sediment particles and porewater (Lohrer et  
43 al., 2004, 2005; Kristensen et al., 2012), and indirectly, by suppressing the presence of tube  
44 building polychaetes and small bivalve species (authors unpublished observation).

45 One well-studied species of the genus *Echinocardium*, *E. cordatum* (Pennant, 1777), forages  
46 by burying itself at depths of up to 20 cm in sandy sediment and up to 6 cm in muddy (silt  
47 and clay) cohesive sediment (Buchanan, 1966; Foster-Smith, 1978; De Ridder & Saucède,  
48 2020). Once buried, *E. cordatum* ploughs the sediment horizontally at speeds ranging from  
49 1–8 cm h<sup>-1</sup> (Buchanan, 1966; Lohrer et al., 2005; Vopel et al., 2007), selectively consuming  
50 microalgae and organic detritus (Smith, 1980; Boon & Duineveld, 2012; De Ridder et al.,  
51 2020). When foraging, *E. cordatum* can displace and mix 60–150 times more sediment  
52 particles than the amount it ingests (Hollertz & Duchêne, 2001; Lohrer et al., 2005).

53 To meet its respiratory requirements while buried, *E. cordatum* maintains a current that  
54 irrigates the restructured sediment with oxygenated seawater, thereby enhancing the  
55 oxidation of reduced solutes and particulates, a process that would otherwise be

56 constrained by the slow molecular diffusion of oxygen through cohesive sediment. Using  
57 dorsal spines and tube feet, the urchin constructs and sustains a vertical funnel, connecting  
58 the ambulacrum with the sediment surface (Buchanan, 1966; Smith, 1980). The cilia of the  
59 ambulacral groove move oxygenated seawater down this funnel, facilitating gas exchange  
60 over respiratory tube feet (Buchanan, 1966; De Ridder & Saucède, 2020). To prevent  
61 suffocation due to the collapse of the respiratory funnel, the funnel-building tube feet  
62 excrete and apply mucus to the wall of the funnel (Kanazawa, 1995; De Ridder & Saucède,  
63 2020). Once the urchin has abandoned the funnel, microbial activity degrades this mucus,  
64 causing the funnel to eventually collapse (Foster-Smith, 1978). Individuals dwelling in  
65 shallow cohesive sediment often maintain a second, larger funnel that connects the  
66 sediment surface in front of the urchin with the urchin's mouth (Vopel et al. 2007).  
67 Furthermore, subanal spines and tube feet construct and sustain a sanitary drain, extending  
68 laterally behind the urchin to accommodate excrements (De Ridder & Saucède, 2020).

69 The displacement of particles and porewater by *E. cordatum* may create effects that last  
70 beyond the duration of the displacement event. For example, the tracks of *E. cordatum*  
71 contribute to the small-scale roughness of the sediment surface, potentially influencing the  
72 solute transport across the sediment–seawater interface (Buchanan, 1966. Vopel et al.  
73 2007). Furthermore, particle mixing can stimulate sedimentary remineralisation and  
74 oxidation processes by, respectively, introducing fresh organic matter into subsurface  
75 sediment layers either in the form of excrements or by displacement from the surface, and  
76 displacing reduced particles at depth to the surface, where they are exposed to oxygenated  
77 seawater (Aller, 1982; Glud, 2008; Lohrer et al., 2004; Vopel et al., 2007, Braeckman et al.  
78 2010). These effects can collectively enhance the remineralisation of labile and refractory  
79 organic carbon, potentially altering the rates of the sediment–seawater exchange of  
80 dissolved carbon and nitrogen.

81 Here, we asked if the previously demonstrated influence of moving *E. cordatum* on  
82 sedimentary solute transport and reaction processes extends beyond the sediment  
83 immediately next to the urchin. In other words, we asked if moving *E. cordatum* cause a  
84 'legacy effect' that sustains a sediment–seawater solute exchange different from the  
85 exchange across a similar but never disturbed sediment? If so, then removing *E. cordatum*  
86 from the seafloor ecosystem should result in measurable difference between the solute

87 exchange across the now undisturbed surface sediment and the surface of the repeatedly  
88 reworked sediment.

89 We experimentally tested this prediction by removing urchins from eight plots of the 10-m  
90 deep seafloor of Man O'War Bay, Hauraki Gulf, New Zealand, and then preventing  
91 recolonisation and thus sediment reworking for about two months. Following this  
92 manipulation, we measured the sediment–seawater exchange of  $O_2$ ,  $NO_3^-$ ,  $NO_2^-$ ,  $NH_4^+$ , and  
93  $N_2$  within and outside the exclusion areas. Because this exchange may be influenced by the  
94 activity of benthic algae and cyanobacteria, we conducted these measurements under  
95 conditions of light and darkness.

## 96 Methods and Materials

### 97 *Study site*

98 Man O'War Bay is situated on the east coast of Waiheke Island in the Hauraki Gulf, New  
99 Zealand (S 36° 47' 38", E 175° 10' 14", Fig. 1). The Hauraki Gulf covers nearly 4,000 km<sup>2</sup>, with  
100 water depths up to 50 m and provides sheltered conditions that facilitate particle settling.  
101 The sediment in Man O'War Bay, which consists of terrigenous silt and clay with a small  
102 proportion of sand, is iron-rich, resulting in a dark grey-to-black subsurface layer due to an  
103 abundance of iron sulphides. Organic matter accounts for >6 % of the sediment dry weight,  
104 but the pore water of the top 20 cm of the sediment remains free of dissolved sulphides  
105 (Wilson & Vopel 2012). In May 2022, the salinity and temperature of the bottom seawater in  
106 Man O'War Bay were 34.5 and ~15 °C, respectively.

107 The seafloor at our 10-m deep site in Man O'War Bay is devoid of macroalgae and sessile  
108 epifauna, with golden-brown diatom patches that accentuate the centimetre-scale surface  
109 topography shaped by two bioturbators: a burrowing mud shrimp of the genus *Upogebia*,  
110 and the ploughing heart urchin *Echinocardium cordatum* (Fig. 2A). Other macroinfauna  
111 consists of small polychaetes (*Prionospio*, *Sthenelais*, and *Cossura*), an amphipod species  
112 (*Paraphoxus*), and a bivalve species (*Theora lubrica*) (Wong & O'Shea 2010).

113 To describe the resident population of *E. cordatum*, a seafloor survey was conducted two  
114 weeks before the exclusion experiment. SCUBA divers used a GoPro Hero 4 camera to  
115 capture a still images of a 0.5 × 0.5-meter quadrat placed on the surface of the seafloor at  
116 40 locations randomly distributed within a 1260 m<sup>2</sup> area. At 10 locations, time-lapse

117 photography was used to estimate the speed at which individuals ploughed the sediment.  
118 The frame that held the underwater camera (Fig. 2A) was positioned within the camera's  
119 field of view, providing a reference for measuring distance.

### 120 *Experimental procedures*

121 On 3 May 2022, two SCUBA divers buried nine plastic rings (1 m diameter, 28 cm wide)  
122 within a ~700 m<sup>2</sup> circular area. The placement of each ring was carried out with a 30-m  
123 guide rope anchored in the sediment and nine pre-determined compass/distance  
124 coordinates (i.e., 0–360 degrees, 1–30 m away from the anchor). The divers pushed each  
125 ring ~25 cm deep into the sediment, leaving about 3 cm exposed above the sediment  
126 surface. Subsequently, they carefully removed all urchins (10–20 individuals per ring) from  
127 the area within the rings. The rings remained free of *E. cordatum* and in place for 56 days,  
128 effectively isolating the enclosed sediment from further sediment restructuring by the  
129 urchins.

130 On 27 June 2022, the divers returned to collect one intact sediment core from within  
131 (Treatment) and outside (Control) each of nine rings, at least 20 cm away from the edge of a  
132 ring, avoiding the sediment around openings of shrimp burrows shown in Figure 2A. To do  
133 so, they placed an acrylic tube (height: 300 mm, internal diameter: 90 mm) perpendicular to  
134 the seafloor and pushed 200 mm of the tube into the sediment until two-thirds were filled.  
135 They then inserted a lid with an open valve into the protruding end of the tube. This was  
136 done slowly, to avoid creating pressure inside the tube that would otherwise resuspend the  
137 silt sediment. Following this, the divers inserted a lid in the buried end of the tube pushing  
138 the sediment core inside the tube upwards by about one centimetre. They then closed the  
139 valve in the top lid and lifted the core out of the sediment. The sediment cores were stored  
140 upright in two large 50 L boxes containing ice, and transported to the laboratory within two  
141 hours.

### 142 *Laboratory setup*

143 In the laboratory, we removed the top lids of the acrylic tubes, and then submerged the  
144 sediment cores in the recirculating seawater (salinity = 34.5) of two 450-L holding tanks for a  
145 3-day acclimatisation. A third and identical tank (measurement tank) was set up to which

146 sets of four or five cores were transferred for sediment–seawater solute flux measurements  
147 (see below).

148 Each tank was fitted with a pump (3260, Eheim) that moved ~9 L seawater per minute  
149 through a chiller unit (HC Chiller 300A, Hailea) and a UV steriliser (Pond One UV-C 9W,  
150 ClearTec) to a 200 L header tank from which the seawater returned to the tank by gravity.  
151 The chiller maintained the seawater temperature close to the in-situ seawater temperature  
152 ( $15 \pm 0.5$  °C). An additional pump pushed seawater through a UV particle filter (Aqua One  
153 Ocellaris 1400 UVC). A wave maker (SW, Jebao) attached to the inner wall of each tank,  
154 along with two jets of returning seawater from the UV particle filter and the header tank,  
155 ensured that advection was sufficient to prevent the seawater overlying the submerged  
156 sediment cores from becoming stagnant. Each tank had a Kessil A160WE Tuna Blue LED  
157 mounted above it, gradually increasing the intensity of photosynthetically active radiation  
158 (PAR) from 06:00 h to a midday maximum of  $\sim 120 \mu\text{mol quanta m}^{-2} \text{s}^{-1}$  and then gradually  
159 decreased this intensity until 18:00 h, when the LED was turned off. To account for  
160 evaporation, we monitored and adjusted the salinity of tank water each morning using  
161 reverse osmosis water.

### 162 *Solute flux measurements*

163 We transferred four or five cores at a time from the holding tanks into the measurement  
164 tank to determine the sediment–seawater solute flux under conditions of light and darkness.  
165 The light and dark solute flux was derived from the difference in the solute concentration in  
166 seawater collected ~10 mm above the sediment surface before and after a 4-h incubation  
167 period. This seawater was withdrawn slowly, using a tube connected to a syringe, avoiding  
168 particle resuspension. During the 4-h incubation period, the acrylic tube holding the  
169 sediment core and its overlying seawater was closed with a valved O-ring sealed lid. A  
170 peristaltic pump recirculated the enclosed seawater through Teflon tubes at a flow rate of  
171  $135 \text{ mL min}^{-1}$  to prevent stagnation. The seawater  $\text{O}_2$  concentration was measured at the  
172 beginning and the end of each incubation with a dipping probe (DP-PSt3, Presens GmbH)  
173 connected to a portable fibre optic oxygen meter (Microx 4, Presens GmbH). The  $\text{O}_2$   
174 concentration in the enclosed seawater never decreased below 86% saturation.

## 175 *Sample analyses*

176 For analyses of nitrite-N, nitrate+nitrite-N, and ammonia-N we collected, filtered (Sartorius  
177 CA 0.45  $\mu\text{m}$  filter), and froze 10 mL of the seawater from each core, at the beginning and the  
178 end of each incubation. These samples were kept frozen until analyses. Seawater aliquots  
179 collected for  $\text{N}_2$  analyses ( $2 \times 12$  mL exetainers) were poisoned with 0.01 mL of a saturated  
180 mercuric chloride solution. Following the seawater sampling, we decanted the seawater  
181 above the sediment cores, extruded the cores, and extracted the top 2 cm of sediment for  
182 granulometric analyses.

183 We determined the sediment granulometric indices by laser particle size analyses  
184 (Mastersizer 2000, Malvern Instruments Ltd.) of 10 homogenised samples, one taken from  
185 the top 2 cm of each of 5 Control and 5 Treatment cores. The sediment water content was  
186 determined as the loss of weight after 24 h drying at 90 °C. The seawater  $\text{N}_2$  content was  
187 measured with a Pfeiffer PrismaPlus QME mass spectrometer and a Bay Instruments S-25-  
188 75D membrane inlet, and the seawater concentrations of nitrite-N, nitrate+nitrite-N, and  
189 ammonia-N were determined with an Astoria Pacific 2 micro-segmented flow analyser  
190 (<https://astoria-pacific.com/>) following the Astoria protocols A182-A00 (range: 0.02–2.0  
191  $\mu\text{mol L}^{-1}$ ), A177-A00 (range: 0.05–7.5  $\mu\text{mol L}^{-1}$ ) and A027-A00 (range: 0.05–5  $\mu\text{mol L}^{-1}$ ) for  
192 operating their respective individual nutrient channels (see also Rho et al. 2015).

## 193 *Solute flux estimates*

194 We determined the sediment–seawater solute flux ( $\mu\text{mol m}^{-2} \text{h}^{-1}$ ) multiplying the start–end  
195 difference in solute concentrations ( $\mu\text{mol L}^{-1}$ ) by the volume of enclosed seawater (L), and  
196 dividing this by the surface area of the sediment core ( $\text{m}^2$ ) and the duration of the  
197 incubation (h). Because previous measurements revealed that the consumption/production  
198 of solutes in the seawater itself is negligible, this was not considered in our flux calculation.  
199 Note that a positive flux indicates sediment consumption whilst a negative flux is interpreted  
200 as solute release from the sediment.

## 201 *Statistical Analysis*

202 All statistical data analyses were completed with R-4.4.0 (R Core Team, 2022). For each  
203 solute we used Shapiro-Wilks test to confirm a normal distribution of the continuous  
204 variable. We then performed a two-way analysis of variance (ANOVA) investigating the

205 effects of light (light versus dark conditions), experimental treatment (cores from within the  
206 exclusion rings versus cores from outside the rings), and the interaction between these  
207 discrete predictors, on the direction and magnitude of each solute flux. Visual analysis of the  
208 diagnostic plots for each of the five solutes confirmed that they met the assumptions of an  
209 ANOVA, including homogeneity of variance. When a significant result was found (alpha  
210 threshold = 0.05), we followed with post-hoc TukeyHSD analyses to further investigate  
211 significant individual effects.

## 212 Results

### 213 *Habitat*

214 The sediment in the sheltered Man O'War Bay consisted of ~74% silt, 17% sand, and ~9%  
215 clay. The laser particle size analyses returned the following granulometric indices: Median =  
216 30  $\mu\text{m}$  (silt), Lower Quartile Q1 = 63  $\mu\text{m}$ , Upper Quartile Q3 = 10  $\mu\text{m}$ , Inclusive Sorting  
217 Coefficient QD1 = 1.95 (poorly sorted), and Inclusive Graphic Skewness  $Sk_1 = 0.07$ . The water  
218 content of the sediment decreased from 75% in the top 2 cm to 65% at 9 cm depth.

219 Analyses of 100 photographs of the sediment surface revealed  $\sim 18.2 \pm 6.7$  (mean  $\pm$  SD, n =  
220 100) *Echinocardium cordatum* individuals per square meter. Time-lapse videos showed that  
221 urchins moved at  $\sim 45 \pm 12$  cm per day (mean  $\pm$  SD, n = 6).

### 222 *Experimental control*

223 The sediment outside the exclusion rings, between foraging *E. cordatum*, and away from the  
224 immediate surrounding of the urchin, consumed  $\text{O}_2$  at similar rates under conditions of  
225 darkness and light (Table 1, Figs 2B, 3). In light, this sediment took up on average ~39% more  
226  $\text{N}_2$  from the overlying seawater than under conditions of darkness (Table 1, Figs 2B, 3). The  
227 average sediment–seawater fluxes of nitrate and nitrite were an order of magnitude lower  
228 than that of  $\text{N}_2$ ; on average nitrate was released (negative flux in Table 1, Figs 2B, 3) while  
229 nitrite was taken up by the sediment, at similar rates under conditions of light and darkness  
230 (Figs 2B, 3).

231 The flux of ammonium in light differed significantly from that measured in darkness ( $p <$   
232 0.025, Table 1, 2). In light, the sediment released ammonium into the overlying seawater

233 (negative flux in Table 1), while the average of fluxes measured in darkness indicated a small  
234 uptake (Table 1, Figs 2B, 3).

### 235 *Sea urchin exclusion—experimental treatment*

236 The solute fluxes ( $O_2$ ,  $N_2$ ,  $NO_3^-$ ,  $NO_2^-$ ,  $NH_4^+$ ) across the surface of the sediment collected  
237 within the exclusion rings (Treatment) did not significantly differ from the fluxes measured  
238 across the surface of the sediment collected outside the exclusion rings (Control), under  
239 conditions of both light and darkness (Tables 1, 2, Fig. 3).

## 240 Discussion

241 We expected that the exchange of dissolved oxygen and nitrogen across the surface of the  
242 sediment that lies beyond the immediate surrounding of foraging *E. cordatum* individuals  
243 would reflect a history of repeated reworking, so that after a 56-day cessation of this  
244 reworking the solute fluxes measured inside (Treatment) and outside (Control) the urchin  
245 exclusion rings would differ in their magnitude and/or direction. Our measurements,  
246 however, did not reveal any statistically significant differences between Treatment and  
247 Control. We considered the potential influence of *E. cordatum* on sedimentary solute  
248 reaction processes; repeated particle mixing and porewater oxygenation should have altered  
249 remineralisation rates and thus modified the associated sediment–seawater solute  
250 exchange. Apparently, the rate at which the sediment was mixed and oxygenated by the  
251 resident *E. cordatum* population did not cause the expected influence, so that the removal  
252 of urchins didn't make a difference.

253 Besides the potential direct effects of sediment mixing on microbial reaction processes, we  
254 expected indirect effects, which can result from the influence of *E. cordatum* on cohabiting  
255 biota, most importantly, benthic microphytes. Photosynthesis of microphytes can cause  
256 dark/light differences in the total sediment–seawater  $O_2$  exchange as measured by core  
257 incubations, and an influence of *E. cordatum* on the abundance and activity of microphytes  
258 would then be evident from either an increase or decrease of these differences. Our  
259 sediment core incubations revealed, however, that the Control removed as much  $O_2$  from  
260 the bottom seawater in darkness than it did in light (Table 1). Assuming that the  $O_2$  demand  
261 of the sediment in light did not differ from that in darkness, this suggests that benthic  
262 photosynthetic  $O_2$  production must have been negligible. If this was the result of heart

263 urchin predation limiting the growth of microphytes, then we would expect algae growth  
264 and photosynthesis to increase following the exclusion of the heart urchins. However, there  
265 was no evidence of such increase; apparently, benthic primary production in the Control  
266 sediment was not limited by predation, so that removal of predation as a potential factor  
267 made no difference.

268 Other studies reported a positive correlation between the abundance of *E. cordatum*  
269 (consuming microphytobenthos) and the chlorophyll-*a* content of its surrounding sediment  
270 (Needham et al., 2011; Lohrer et al., 2004, 2005, 2015). In habitats where photosynthetically  
271 active radiation (PAR) is not limiting microphytobenthic production, this may follow from a  
272 stimulation of bacterial mineralisation processes, increasing the porewater concentrations of  
273 inorganic nitrogen, so that algal growth outweighs consumption. Such gardening effect will  
274 be subject to seasonal variations in environmental conditions and heart urchin abundance  
275 and activity. Lower PAR and temperature in winter, for example, can hamper benthic primary  
276 production and slow urchin movement, respectively (Seike et al., 2022). A repeat of our  
277 study in November/December may reveal if the influence of *E. cordatum* on the sediment–  
278 seawater solute exchange at our site in fact depends on seasonal changes in conditions for  
279 growth and production of microphytes or if what we have observed in May/June generally  
280 applies throughout the year.

281 Interestingly, both sediments, Control and Treatment, served as a sink for N<sub>2</sub>. Others have  
282 reported nitrogen fixation in ammonium-rich sediments, like those found in seagrass beds  
283 (Aoki & McGlathery, 2019) or subtidal coastal sediments (Fulweiler et al., 2007, Newell et al.,  
284 2016), and such fixation may be limited to microenvironments in which nitrogen becomes  
285 limiting (Whiting et al. 1986). As Bertics et al. (2013) pointed out, however, sediments rich in  
286 organic matter and reduced substrates provide an unusual conundrum for anaerobic  
287 organisms—the excess electrons can damage cells and there is a need to regenerate  
288 electron carriers. Nitrogen fixation may be a mechanism for dissipating reducing power and  
289 maintaining an ideal intracellular redox state (Joshi & Tabita, 1996, Bombar et al., 2016), so  
290 that nitrogen fixation is observed despite high ammonium concentrations.

291 The significant light–dark difference in the ammonium exchange of the Control sediment—a  
292 small uptake under conditions of darkness, but release in light—suggests an influence of  
293 photosynthesis on nitrogen cycling. This is contrast to the interpretation of our light/dark O<sub>2</sub>

294 exchange data (negligible benthic photosynthesis, see above), and raises the question if our  
295 assumption of similar sediment O<sub>2</sub> demand in light and darkness was valid. If the sediment  
296 O<sub>2</sub> demand in light was greater than in darkness, then benthic photosynthesis may have  
297 played a role without leaving a measurable signature in the total sediment–seawater O<sub>2</sub>  
298 exchange. Assuming that the ammonium and N<sub>2</sub> concentrations in the sediment overlying  
299 seawater remained stable and independent of light, then the difference in the derived  
300 sediment–seawater flux must have resulted from changes in the sediment porewater solute  
301 concentrations. Phototrophs assimilating ammonium in darkness, but not in light, would  
302 then cause a light/darkness difference in the sediment–seawater ammonium exchange.  
303 Similarly, if more porewater N<sub>2</sub> was to be consumed by diazotrophs in light than in darkness  
304 (energy limited N<sub>2</sub> fixation; Stal, 2015), then this would explain why the sediment uptake  
305 was greater in light than in darkness. Alternatively, a production of N<sub>2</sub> via anammox in  
306 darkness would increase the porewater N<sub>2</sub> concentration and so lower the rate at which the  
307 sediment removes N<sub>2</sub> from the bottom seawater. Because this process also lowers the  
308 porewater ammonium concentration, it can explain both observations, a lower release or  
309 even uptake of ammonium, and lower uptake of N<sub>2</sub> under conditions of darkness.

310 In conclusion, although foraging *E. cordatum* enhance the sediment–seawater solute  
311 exchange where and when they displace sediment particles, this effect does not necessarily  
312 outlast the disturbance event. At our subtidal site, at least in May/June, removal of the heart  
313 urchins did not alter the solute exchange across the surface of the sediment area that is not  
314 actively disturbed. This underlines the importance of evaluating the influence of  
315 bioturbators on ecosystem processes in the context of local environmental conditions,  
316 seasonal changes in these conditions and biological activity, urchin behaviour, and  
317 population characteristics.

## 318 Acknowledgments

319 Evan Brown assisted in the field. K.V. designed the experiment, R.M. and M.S. with the  
320 assistance of Mohammad H. A. Usmani completed the laboratory sediment incubations.  
321 M.S. measured the inorganic nitrogen content of seawater samples, and R.M. and K.V.  
322 analysed the data and wrote the manuscript.

323

## 324 References

- 325 Aller, R. C. (1982). The Effects of macrobenthos on chemical properties of marine sediment  
326 and overlying water. In P. L. McCall & M. J. S. Tevesz (Eds.), *Animal-Sediment*  
327 *Relations* (Vol. 100, pp. 53–102). Springer US. [https://doi.org/10.1007/978-1-4757-](https://doi.org/10.1007/978-1-4757-1317-6_2)  
328 [1317-6\\_2](https://doi.org/10.1007/978-1-4757-1317-6_2)
- 329 Aller, R. C., & Aller, J. Y. (1998). The effect of biogenic irrigation intensity and solute exchange  
330 on diagenetic reaction rates in marine sediments. *Journal of Marine Research*, 56(4),  
331 905–936. <https://doi.org/10.1357/002224098321667413>
- 332 Aoki, L., & McGlathery, K. (2019). High rates of N fixation in seagrass sediments measured via  
333 a direct  $^{30}\text{N}_2$  push-pull method. *Marine Ecology Progress Series*, 616, 1–11.  
334 <https://doi.org/10.3354/meps12961>
- 335 Banta, G., Holmer, M., Jensen, M., & Kristensen, E. (1999). Effects of two polychaete worms,  
336 *Nereis diversicolor* and *Arenicola marina*, on aerobic and anaerobic decomposition in  
337 a sandy marine sediment. *Aquatic Microbial Ecology*, 19, 189–204.  
338 <https://doi.org/10.3354/ame019189>
- 339 Bertics, V. J., Löscher, C. R., Salonen, I., Dale, A. W., Gier, J., Schmitz, R. A., & Treude, T.  
340 (2013). Occurrence of benthic microbial nitrogen fixation coupled to sulfate  
341 reduction in the seasonally hypoxic Eckernförde Bay, Baltic Sea. *Biogeosciences*,  
342 10(3), 1243–1258. <https://doi.org/10.5194/bg-10-1243-2013>
- 343 Bombar, D., Paerl, R. W., & Riemann, L. (2016). Marine non-cyanobacterial diazotrophs:  
344 Moving beyond molecular detection. *Trends in Microbiology*, 24(11), 916–927.  
345 <https://doi.org/10.1016/j.tim.2016.07.002>
- 346 Boon, A. R., & Duineveld, G. C. A. (2012). Phytopigments and fatty acids in the gut of the  
347 deposit-feeding heart urchin *Echinocardium cordatum* in the southern North Sea:  
348 Selective feeding and its contribution to the benthic carbon budget. *Journal of Sea*  
349 *Research*, 67(1), 77–84. <https://doi.org/10.1016/j.seares.2011.10.004>
- 350 Braeckman, U., Provoost, P., Gribsholt, B., Van Gansbeke, D., Middelburg, J.J., Soetaert, K.,  
351 Vincx, M. and Vanaverbeke, J. (2010). Role of macrofauna functional traits and  
352 density in biogeochemical fluxes and bioturbation. *Marine Ecology Progress Series*,  
353 399, 173–186.

354 Buchanan, J. B. (1966). The biology of *Echinocardium cordatum* [Echinodermata:  
355 Spatangoidea] from different habitats. *Journal of the Marine Biological Association of*  
356 *the United Kingdom*, 46(1), 97–114. <https://doi.org/10.1017/S0025315400017574>

357 De Ridder, C., Jangoux, M., & Impe, E. (2020). Food selection and absorption efficiency in the  
358 spatangoid echinoid, *Echinocardium cordatum* (Echinodermata). In B. F. Keegan & B.  
359 D. S. O'Connor (Eds.), *Echinodermata* (1st ed., pp. 245–251). CRC Press.  
360 <https://doi.org/10.1201/9781003079224-48>

361 De Ridder, C., & Saucède, T. (2020). *Echinocardium cordatum*. In *Developments in*  
362 *Aquaculture and Fisheries Science* (Vol. 43, pp. 337–357). Elsevier.  
363 <https://doi.org/10.1016/B978-0-12-819570-3.00020-2>

364 Forster, S., Graf, G., Kitlar, J., & Powilleit, M. (1995). Effects of bioturbation in oxic and  
365 hypoxic conditions: A microcosm experiment with a North Sea sediment community.  
366 *Marine Ecology Progress Series*, 116, 153–161. <https://doi.org/10.3354/meps116153>

367 Foster-Smith, R. L. (1978). An analysis of water flow in tube-living animals. *Journal of*  
368 *Experimental Marine Biology and Ecology*, 34(1), 73–95.  
369 [https://doi.org/10.1016/0022-0981\(78\)90058-8](https://doi.org/10.1016/0022-0981(78)90058-8)

370 Fulweiler, R. W., Nixon, S. W., Buckley, B. A., & Granger, S. L. (2007). Reversal of the net  
371 dinitrogen gas flux in coastal marine sediments. *Nature*, 448(7150), 180–182.  
372 <https://doi.org/10.1038/nature05963>

373 Glud, R. N. (2008). Oxygen dynamics of marine sediments. *Marine Biology Research*, 4(4),  
374 243–289. <https://doi.org/10.1080/17451000801888726>

375 Hollertz, K., & Duchêne, J. C. (2001). Burrowing behaviour and sediment reworking in the  
376 heart urchin *Brissopsis lyrifera* Forbes (Spatangoida). *Marine Biology*, 139(5), 951–  
377 957. <https://doi.org/10.1007/s002270100629>

378 Joshi, H. M., & Tabita, F. R. (1996). A global two component signal transduction system that  
379 integrates the control of photosynthesis, carbon dioxide assimilation, and nitrogen  
380 fixation. *Proceedings of the National Academy of Sciences*, 93(25), 14515–14520.  
381 <https://doi.org/10.1073/pnas.93.25.14515>

382 Kanazawa, K. (1995). How spatangoids produce their traces: Relationship between  
383 burrowing mechanism and trace structure. *Lethaia*, 28(3), 211–219.  
384 <https://doi.org/10.1111/j.1502-3931.1995.tb01424.x>

385 Kristensen, E., Penha-Lopes, G., Delefosse, M., Valdemarsen, T., Quintana, C., & Banta, G.  
386 (2012). What is bioturbation? The need for a precise definition for fauna in aquatic  
387 sciences. *Marine Ecology Progress Series*, 446, 285–302.  
388 <https://doi.org/10.3354/meps09506>

389 Lohrer, A. M., Thrush, S. F., & Gibbs, M. M. (2004). Bioturbators enhance ecosystem function  
390 through complex biogeochemical interactions. *Nature*, 431(7012), 1092–1095.  
391 <https://doi.org/10.1038/nature03042>

392 Lohrer, A. M., Thrush, S. F., Hewitt, J. E., & Kraan, C. (2015). The up-scaling of ecosystem  
393 functions in a heterogeneous world. *Scientific Reports*, 5(1), 10349.  
394 <https://doi.org/10.1038/srep10349>

395 Lohrer, A. M., Thrush, S. F., Hunt, L., Hancock, N., & Lundquist, C. (2005). Rapid reworking of  
396 subtidal sediments by burrowing spatangoid urchins. *Journal of Experimental Marine  
397 Biology and Ecology*, 321(2), 155–169. <https://doi.org/10.1016/j.jembe.2005.02.002>

398 Nakamura, Y. (2001). Autoecology of the heart urchin, *Echinocardium cordatum*, in the  
399 muddy sediment of the Seto Inland Sea, Japan. *Journal of the Marine Biological  
400 Association of the United Kingdom*, 81(2), 289–297.  
401 <https://doi.org/10.1017/S0025315401003769>

402 Nedwell, D. B., & Walker, T. R. (1995). Sediment-water fluxes of nutrients in an Antarctic  
403 coastal environment: Influence of bioturbation. *Polar Biology*, 15(1).  
404 <https://doi.org/10.1007/BF00236125>

405 Needham, H. R., Pilditch, C. A., Lohrer, A. M., & Thrush, S. F. (2011). Context-specific  
406 bioturbation mediates changes to ecosystem functioning. *Ecosystems*, 14(7), 1096–  
407 1109. <https://doi.org/10.1007/s10021-011-9468-0>

408 Newell, S. E., McCarthy, M. J., Gardner, W. S., & Fulweiler, R. W. (2016). Sediment nitrogen  
409 fixation: A call for re-evaluating coastal N budgets. *Estuaries and Coasts*, 39(6), 1626–  
410 1638. <https://doi.org/10.1007/s12237-016-0116-y>

411 Osinga, R., Kop, A. J., Malschaert, J. F. P., & Van Duyl, F. C. (1997). Effects of the sea urchin  
412 *Echinocardium cordatum* on bacterial production and carbon flow in experimental  
413 benthic systems under increasing organic loading. *Journal of Sea Research*, 37(1–2),  
414 109–121. [https://doi.org/10.1016/S1385-1101\(97\)00003-8](https://doi.org/10.1016/S1385-1101(97)00003-8)

415 R Core Team. (2022). R: A language and environment for statistical computing. [Computer  
416 software]. R Foundation for Statistical Computing. <https://www.R-project.org/>

417 Rho, T, Coverly, S., Kim, E.-S., Kang, D.-J., Kahn, S.-H., Na, T.-H., Cho, S.-R., Lee, J.-M., Moon, C.-  
418 R. (2016) Practical considerations for the segmented-flow analysis of nitrate and  
419 ammonium in seawater and the avoidance of matrix effects. *Ocean Science Journal*,  
420 50(4), 709–720. <https://doi.org/10.1007/s12601-015-0064-7>

421 Seike, K., Sassa, S., Shirai, K., & Kubota, K. (2022). Sediment hardness and water temperature  
422 affect the burrowing of *Echinocardium cordatum*: Implications for mass mortality  
423 during the 2011 earthquake–liquefaction–tsunami disaster. *Estuarine, Coastal and  
424 Shelf Science*, 267, 107763. <https://doi.org/10.1016/j.ecss.2022.107763>

425 Smith, A. B. (1980). The structure and arrangement of echinoid tubercles. *Philosophical  
426 Transactions of the Royal Society of London. B, Biological Sciences*, 289(1033), 1–54.  
427 <https://doi.org/10.1098/rstb.1980.0026>

428 Solan, M., & Wigham, B. D. (2005). Biogenic particle reworking and bacterial–invertebrate  
429 interactions in marine sediments. In E. Kristensen, R. R. Haese, & J. E. Kostka (Eds.),  
430 *Coastal and Estuarine Studies* (Vol. 60, pp. 105–124). American Geophysical Union.  
431 <https://doi.org/10.1029/CE060p0105>

432 Stal, L. J. (2015). Nitrogen Fixation in Cyanobacteria. In *Wiley, Encyclopedia of Life Sciences*  
433 (1st ed., pp. 1–9). Wiley. <https://doi.org/10.1002/9780470015902.a0021159.pub2>

434 Volkenborn, N., Hedtkamp, S. I. C., van Beusekom, J. E. E., & Reise, K. (2007). Effects of  
435 bioturbation and bioirrigation by lugworms (*Arenicola marina*) on physical and  
436 chemical sediment properties and implications for intertidal habitat succession.  
437 *Estuarine, Coastal and Shelf Science*, 74(1–2), 331–343.  
438 <https://doi.org/10.1016/j.ecss.2007.05.001>

439 Vopel, K., Vopel, A., Thistle, D., & Hancock, N. (2007). Effects of spatangoid heart urchins on  
440 O<sub>2</sub> supply into coastal sediment. *Marine Ecology Progress Series*, 333, 161–171.  
441 <https://doi.org/10.3354/meps333161>

442 Whiting, G. J., Gandy, E. L., & Yoch, D. C. (1986). Tight coupling of root-associated nitrogen  
443 fixation and plant photosynthesis in the salt marsh grass *Spartina alterniflora* and  
444 carbon dioxide enhancement of nitrogenase activity. *Applied and Environmental  
445 Microbiology*, 52(1), 108–113. <https://doi.org/10.1128/aem.52.1.108-113.1986>

446 Wilson, P. S., Vopel, K. (2012) Estimating the in situ distribution of acid volatile sulfides from  
447 sediment profile images. *Limnology and Oceanography: Methods*, 10(12), 1070–  
448 1077. <https://doi.org/10.4319/lom.2012.10.1070>

449 Wong, K. C., & O'Shea, S. (2010). Sediment macrobenthos off eastern Waiheke Island,  
450 Hauraki Gulf, New Zealand. *New Zealand Journal of Marine and Freshwater Research*,  
451 44(3), 149–165. <https://doi.org/10.1080/00288330.2010.498088>  
452

453 Figure captions

454 Figure 1. Map depicting the North Island of New Zealand and the study site (insert, yellow  
455 marker) in Man O'War Bay, at the East shore of Waiheke Island, Hauraki Gulf, New  
456 Zealand (S 36° 47' 38", E 175° 10' 14").

457 Figure 2. (A) Sediment surface at Man O'War Bay, Hauraki Gulf, New Zealand, showing tracks  
458 (tr) made by the heart urchin *Echinocardium cordatum*. Orange arrows indicate the  
459 direction of the heart urchin's movement. Black arrows indicate the respiratory (rf) and  
460 feeding (ff) funnels. Larger holes are the openings of shrimp burrows, presumably made  
461 by *Upogebia* sp. Darker areas lining the tracks are diatoms. (B) Schematic illustrating the  
462 direction and magnitude of the sediment–seawater solute flux under conditions of light  
463 (light grey arrows) and darkness (dark grey arrows). The horizontal line indicates the  
464 sediment surface.

465 Figure 3. Sediment–seawater flux ( $\mu\text{mol m}^{-2} \text{h}^{-1}$ ,  $n = 9$ ) of ammonium ( $\text{NH}_4^+$ ), nitrite ( $\text{NO}_2^-$ ),  
466 nitrate ( $\text{NO}_3^-$ ), nitrogen gas ( $\text{N}_2$ ), and oxygen ( $\text{O}_2$ ) in darkness (closed symbols) and light  
467 (open symbols), derived from laboratory incubations of sediment collected in Man O'  
468 War Bay, Hauraki Gulf, New Zealand outside (Control) and inside (Treatment) sea urchin  
469 exclusion rings after 56 days. A positive flux indicates sediment solute uptake. Horizontal  
470 bars are means, the asterisk indicates a significant difference, and the hash sign marks  
471 two groups from each of which one outlier has been removed ( $n = 8$ ).

472

473 Table 1. Mean sediment–seawater flux ( $\mu\text{mol m}^{-2} \text{h}^{-1}$ , Mean  $\pm$  C.I., n = 9) of ammonium  
 474 ( $\text{NH}_4^+$ ), nitrite ( $\text{NO}_2^-$ ), nitrate ( $\text{NO}_3^-$ ), nitrogen gas ( $\text{N}_2$ ), and oxygen ( $\text{O}_2$ ) in darkness and light,  
 475 derived from laboratory incubations of sediment collected in Man O' War Bay, Hauraki Gulf,  
 476 New Zealand outside (Control) and inside (Treatment) sea urchin exclusion rings after 56  
 477 days. A positive flux indicates sediment solute uptake. Number in parenthesis, n; hash sign,  
 478 one outlier removed.

479

	Control		Treatment	
	Darkness	Light	Darkness	Light
$\text{N}_2$	121.4 $\pm$ 28.6 (9)	169.2 $\pm$ 35.1 (9)	93.7 $\pm$ 24.5 (9)	141.3 $\pm$ 30.4 (8) <sup>#</sup>
$\text{O}_2$	149.6 $\pm$ 22.8 (9)	140.4 $\pm$ 23.8 (8) <sup>#</sup>	148.2 $\pm$ 29.0 (9)	137.6 $\pm$ 45.3 (9)
$\text{NO}_2^-$	13.3 $\pm$ 18.9 (9)	8.1 $\pm$ 25.9 (9)	5.8 $\pm$ 7.7 (9)	7.6 $\pm$ 16.2 (9)
$\text{NO}_3^-$	-22.8 $\pm$ 20.8 (9)	-22.9 $\pm$ 26.5 (9)	-18.2 $\pm$ 9.1 (9)	-22.4 $\pm$ 19.3 (9)
$\text{NH}_4^+$	2.5 $\pm$ 29.0 (9)	-49.7 $\pm$ 13.7 (9)	-6.5 $\pm$ 11.3 (9)	-30.4 $\pm$ 11.8 (9)

480

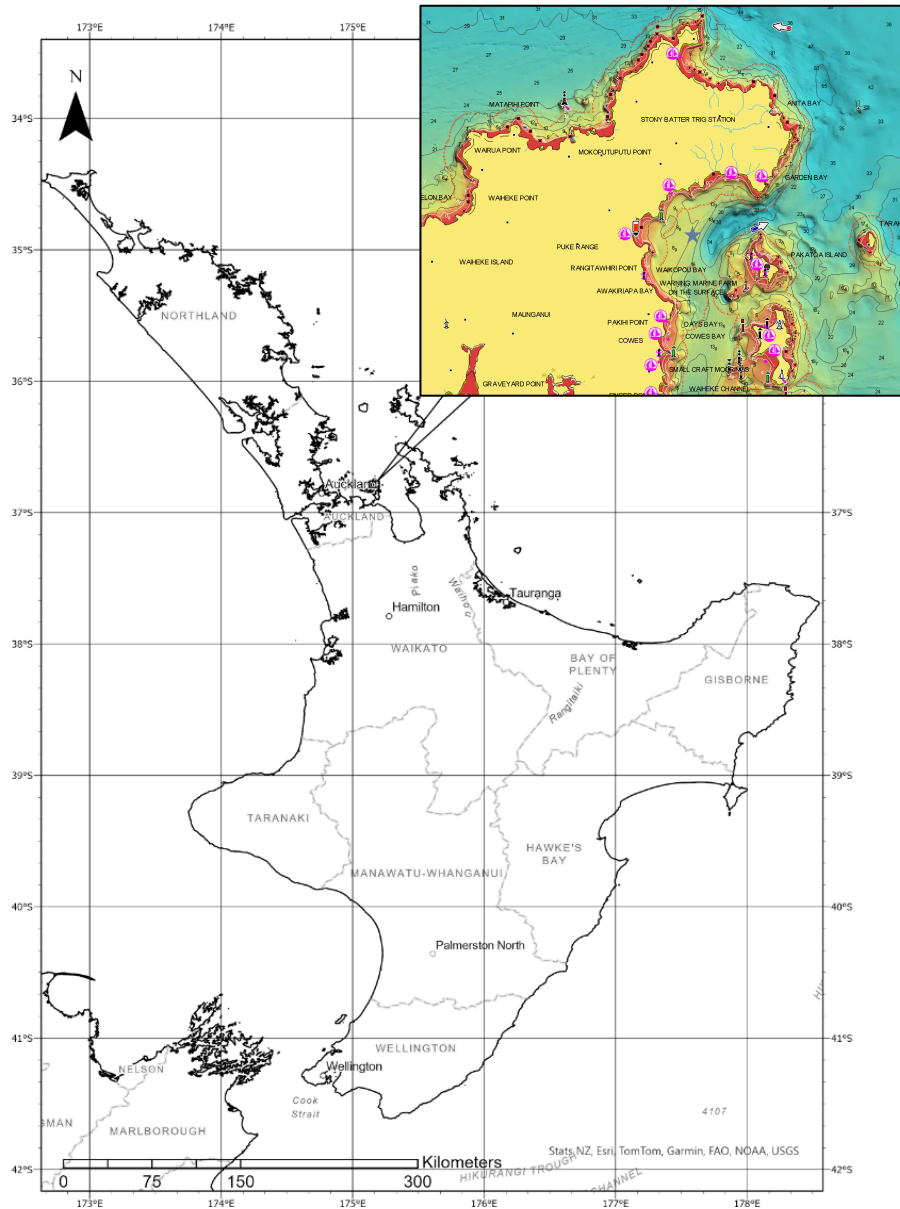
481 Table 2. Two-way ANOVA and TukeyHSD (df, Mean Sum of Squares, F values, adjusted P  
 482 Values, C.I. 95%) of solute flux across the surface of sediment cores as a function of the  
 483 presence (Control) or absence (Treatment) of *Echinocardium cordatum*, the measurement  
 484 light regime (light/dark), and their interaction.

485

Solute	df	Mean SS	F value	P value	TukeyHSD
<b>N<sub>2</sub></b>					
Urchin presence	1	7449	3.658	0.055	
Light regime	1	19835	9.741	<0.004	CL:TD < 0.00066
Interaction	1	0	0	0.996	
<b>O<sub>2</sub></b>					
Urchin presence	1	560	0.164	0.458	
Light regime	1	3707	1.087	0.247	
Interaction	1	395	0.116	0.429	
<b>NO<sub>2</sub><sup>-</sup></b>					
Urchin presence	1	142	0.180	0.674	
Light regime	1	26	0.033	0.857	
Interaction	1	108	0.136	0.714	
<b>NO<sub>3</sub><sup>-</sup></b>					
Urchin presence	1	59	0.063	0.803	
Light regime	1	39	0.042	0.839	
Interaction	1	37	0.040	0.843	
<b>NH<sub>4</sub><sup>+</sup></b>					
Urchin presence	1	240	0.317	0.577	
Light regime	1	13026	17.171	<0.001	CL:CD < 0.025; CL:TD < 0.019
Interaction	1	1804	2.378	0.133	

486

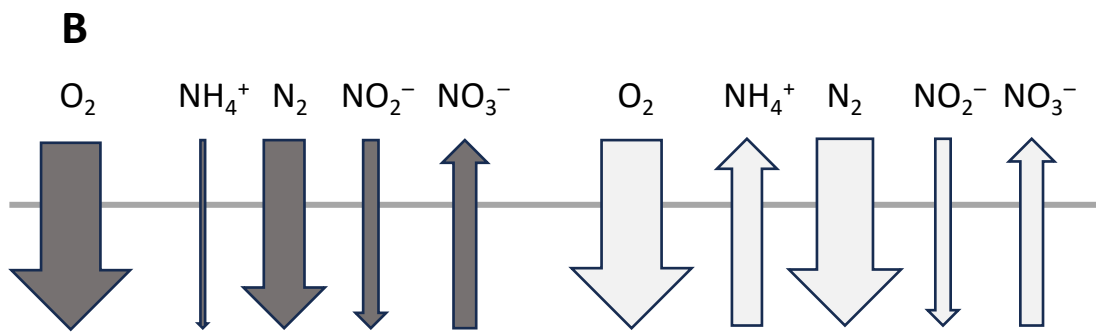
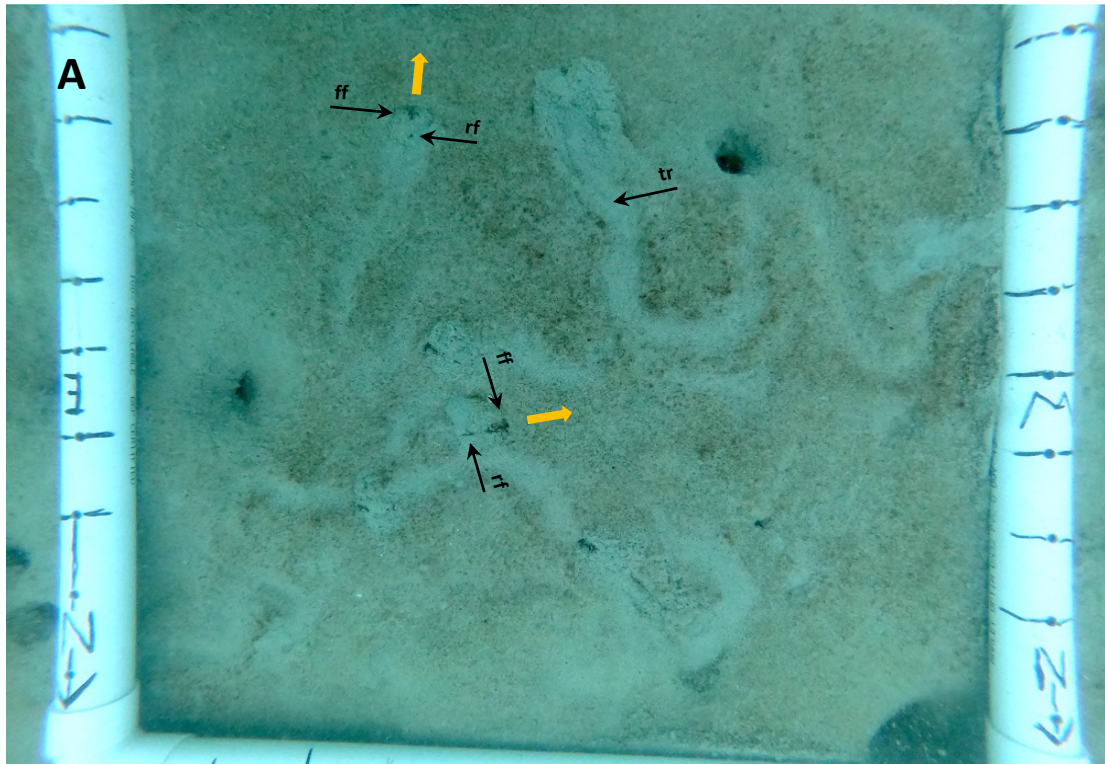
487



488

489 Figure 1.

490



491

492 Figure 2

493

